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Fungi Benefit from Two Decades of Increased Nutrient Availability in Tundra Heath Soil

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Abstract

If microbial degradation of carbon substrates in arctic soil is stimulated by climatic warming, this would be a significant positive feedback on global change. With data from a climate change experiment in Northern Sweden we show that warming and enhanced soil nutrient availability, which is a predicted long-term consequence of climatic warming and mimicked by fertilization, both increase soil microbial biomass. However, while fertilization increased the relative abundance of fungi, warming caused only a minimal shift in the microbial community composition based on the phospholipid fatty acid (PLFA) and neutral lipid fatty acid (NLFA) profiles. The function of the microbial community was also differently affected, as indicated by stable isotope probing of PLFA and NLFA. We demonstrate that two decades of fertilization have favored fungi relative to bacteria, and increased the turnover of complex organic compounds such as vanillin, while warming had no such effects. Furthermore, the NLFA-to-PLFA ratio for \( ^{13} \text{C}\)-incorporation from acetate increased in warmed plots but not in fertilized ones. Thus, fertilization cannot be used as a proxy for effects on warming in arctic tundra soils. Furthermore, the different functional responses suggest that the biomass increase found in both fertilized and warmed plots was mediated via different mechanisms.

Introduction

Microbial use of carbon substrates in soil is of critical importance when estimating impacts of climate change on the fate of soil carbon. Especially in the Arctic, where the amount of soil carbon exceeds the amount of carbon in aboveground plant biomass by an order of magnitude [1], microbial processes transforming the carbon compounds are crucial. In this region, climate change is predicted to increase the mean annual temperature by 2–9°C within this century [2].

Warming generally stimulates microbial growth [3] and thereby increases nutrient mineralization [4] in soil. As enhanced nutrient availability thus is an expected long-term consequence of warming, fertilization is often used as a long-term proxy for warming in experimental studies [5,6]. Both warming and enhanced nutrient availability affect turnover of soil carbon, directly through effects on microbial activity [3] and indirectly by increasing plant growth [7] and altering vegetation composition [8,9].

Changes in plant growth and vegetation composition influence carbon inputs to soil in several ways. In the short term, the quantity and the chemical quality of root exudates may change as these differ among plant species [10]. Also, the distribution of these fresh carbon inputs in the soil horizon may change, as different plant groups have different rooting depths [6,11]. In the long-term, vegetation changes will influence litter input. For example, an increase in deciduous shrubs [12] would lead to higher leaf litter deposition each autumn, while an increase in graminoids [8] would add easily decomposable plant litter to soil [13]. Climate change may also alter the chemical composition of a plant, often increasing the concentration of secondary metabolites in the plant tissue [14].

We aimed to unravel how the composition of microbial communities and their use of carbon substrates in subarctic tundra soil respond to long-term warming and enhanced nutrient availability using SIP-PLFA (stable isotope probing of phospholipid fatty acids). Soil samples were taken from a field experiment which had been running for 18 years prior to the sampling on a mesic/dry heath just above the tree line near Abisko Scientific Research Station in Northern Sweden. The warming treatment, which was accomplished by open-top chambers, initially resulted in an increased biomass of deciduous shrubs [15], decreased abundance of bryophytes [8], and reduced soil microbial biomass [5]. After a decade, warming had thus reduced the frequency of mosses by 50% and that of lichens by 33%, but there was no effect on graminoids [8]. The NPK (nitrogen, phosphorous and potassium) fertilization treatment, which aimed at mimicking increased soil nutrient availability, had also increased vascular plant biomass at the expense of bryophyte biomass [15], but increased soil microbial biomass including fungal biomass [5]. After 14 years, both warming and fertilization had increased the biomass of deciduous shrubs, but the increase in graminoids was
still confined to the fertilized plots [15]. After 22 years, both warming and fertilization treatments had 35% higher vascular plant cover than the control [16].

In the laboratory, the soil samples were amended with a range of 13C-labelled substrates, and the uptake of these substrates into various fatty acids indicating different microbial groups and either growth (PLFAs) or storage (neutral lipid fatty acids, NLFAs) was followed. The selected substrates represented carbon sources present in soil. Glucose, acetic acid and glycine are simple compounds common in plant root exudates, and glycine is also a source of nitrogen for subarctic plants and microbes [11]. Starch is a very common polysaccharide in plant residues. Vanillin is a common product of lignin depolymerisation [17], containing a phenol ring, and is often used as a model substance indicating lignin degradation. Starch and vanillin are therefore examples of more complex substrates and supposedly more difficult to decompose.

We expected that, in line with the previous results from the field experiment [5], fertilization would increase and warming decrease microbial biomass. We hypothesized that fertilization would increase the uptake of carbon substrates into PLFA as compared to NLFA due to alleviated nutrient limitation. Furthermore, we expected that vanillin uptake would decrease with increased N availability, since it has earlier been shown that fertilization can suppress ligninolytic enzymes in forest soil [18–20]. Warming was expected to have the same effects as fertilization if this treatment would lead to increased nutrient availability.

Results

Treatment Effects on Fatty Acid Concentrations

Substrate additions did not cause any significant changes in the total PLFA or NLFA concentration, and therefore an average of each substrate is presented (Fig. 1). Fertilization and warming significantly increased the total PLFA concentration in soil by about 31% and 20%, respectively (Fig. 1; P < 0.001), which indicates that these treatments increased microbial biomass. The total NLFA concentration was only increased by fertilization (P < 0.001), although less when fertilization was combined with warming (Fig. 1; P < 0.01 for warming×fertilization interaction). There were no significant differences between the two incubation times.

In the PCA of the PLFA profiles, the first PC (explained variance 32%) described treatment effects on the PLFA patterns (Fig. 2). Control treatments close to the origin, and the fertilized soil had the highest values (Fig. 2a; P < 0.001), although less when fertilization was combined with warming (Fig. 1; P < 0.01 for warming×fertilization interaction). The first PC accounted for the difference between the fertilized and the unfertilized soils, and showed a warming×fertilization interaction (Fig. 2a; P < 0.001). As for PLFAs, warming×fertilization interaction term was highly significant (P < 0.001) because of the relatively similar PLFA patterns in the warming + fertilization treatments as compared to fertilization alone. The fertilized soil was characterized by relatively higher amounts of the fungal biomarker 18:20:6:9, than the unfertilized soil (Fig. 2b). The second PC (explained variance 15%) mainly accounted for the difference between the incubation times (Fig. 2a; P < 0.001); the relative abundance of the straight-chained PLFAs 14:0, 15:0 and 16:0, and the markers for Gram-negative bacteria i14:0, i15:0, a15:0 and a16:0 increased from one to seven-day samples (Fig. 2b). The fungal biomarker 18:20:6:9 was not affected by the incubation time.

The PCA of the NLFA mole percentage profiles showed that while the 1-day incubation time produced relatively similar patterns except for the fertilization treatment, there was a large spread in the patterns for the 7-day-long incubation (Fig. 3a). As for PLFA, the first PC (explained variance 38%) accounted for the difference between the fertilized and the unfertilized soils, and showed a warming×fertilization interaction (Fig. 3a; P < 0.001). As for PLFAs, the difference was mainly due to a higher abundance of the fungal markers, in this case the NLFAs 18:20:6:9 and 18:16:9 (Fig. 3b). The second PC mainly separated the incubation times from each other (Fig. 3a; P < 0.001).

Treatment Effects on 13C-incorporation into PLFA

There were no significant effects of treatments or incubation time on total 13C-incorporation from glucose, acetic acid, glycine (Table 1). Fertilization increased total 13C-incorporation from vanillin (P < 0.001), but did not significantly affect incorporation from starch. The uptake of starch and vanillin increased from one to seven-day-long incubation (Table 1; P < 0.01).

The PCAs for the 13C-incorporation patterns showed that for all substrate additions, the first two PCs separated the fertilized soil from the unfertilized soil (Fig. 4; P < 0.001). A common explanation for this difference was an increased incorporation of 13C into the fungal biomarker PLFAs, 18:20:6:9 and 18:16:9 and a decreased incorporation into most of the bacterial biomarker PLFAs under fertilization (see Figs S1, S2, S3, S4, S5, S6, S8). As an exception, the Gram-negative bacteria indicated by cy17:0 and cy19:0 took up significantly more glycine, starch and vanillin in the fertilized than in the unfertilized soil (Fig. S3). In addition, utilization of starch was higher in the fertilized than in the unfertilized soil for Gram-positive bacteria containing i16:0 and the actinomycetes indicated by the methylated PLFA 10Me16:0, especially after incubation for seven days (Figs S6 and S7).

The substrate utilization patterns in the warming treatment were rather similar to those in the control samples, except for starch addition with a significant warming effect on the second PC (Fig. 4, P < 0.01). The incorporation of 13C from starch into the PLFAs i16:0 and 10Me16:0 was higher in the warming treatment than in the control soil after incubation for seven days (Figs S6 and S7).
S8). Uptake of glucose by organisms containing the PLFA 18:1\v 9 was significantly reduced by warming (Fig. S2). The warming+fertilization treatment had relatively similar substrate utilization patterns as fertilization alone (Fig. 4), which led to warming\xrightarrow{Fertilization} interactions in the incorporation to some individual PLFAs (Figs S1, S2, S3, S4, S5, S6, S7, S8).

A clear difference in the \textsuperscript{13}C-incorporation patterns between the two incubation times could be observed on the first two PCs for all substrates (Fig. 4; P < 0.001). This owed mainly to the increased incorporation into the PLFAs cy17:0, cy19:0, 10Me16:0 and 10Me18:0 over time (Figs S3, S7, S8).

**Treatment Effects on \textsuperscript{13}C-incorporation into NLFA**

The incorporation of \textsuperscript{13}C to NLFA decreased over time for glucose, acetic acid, and glycine (Table 1; P < 0.01). Fertilization increased the total incorporation of \textsuperscript{13}C from vanillin (P < 0.01) but did not significantly affect incorporation from other substrates.

Incorporation of \textsuperscript{13}C to the fungal biomarker NLFAAs, 18:2\v 6 and 18:1\v 9, was highest from acetic acid, followed by glucose and...
vanillin (Fig. 5a, b). Incorporation of $^{13}$C to these NLFA from glycine and starch was minimal. The response patterns for 18:2ω6,9 and 18:1ω9 were similar, but because of the higher variance, only vanillin-amended samples showed significant differences at $P<0.01$ for 18:2ω6,9; fertilization increased ($P<0.001$) and warming decreased ($P<0.01$) $^{13}$C-incorporation from vanillin (Fig. 5a). Similar responses were observed for $^{13}$C-incorporation from vanillin into the NLFA 18:1ω9 (Fig. 5b; $P<0.001$ for both warming and fertilization). Fertilization also significantly increased $^{13}$C-incorporation from glucose, when not combined with warming treatment ($P<0.01$ for warming×fertilization).

The NLFA-to-PLFA ratios for $^{13}$C-incorporation into 18:2ω6,9 and 18:1ω9 were not significantly affected by treatments for other substrates than acetic acid (data for glucose and acetic acid shown in Table 1). For acetic acid, the ratio for incorporation into 18:2ω6,9 was significantly increased by warming after incubation for 7 days (with a similar insignificant trend for 18:1ω9), while it was unaffected by fertilization (Table 1).
Table 1. Total $^{13}$C-incorporation into phospholipid (PLFA) and neutral lipid fatty acids (NLFA) and NLFA/PLFA ratios for $^{13}$C-incorporation into fungal biomarkers (mean±SE, n=6) in tundra heath soil following incubation with added glucose, glycine, acetic acid, starch or vanillin for 1 or 7 days.

<table>
<thead>
<tr>
<th></th>
<th>1-day incubation</th>
<th>7-day incubation</th>
<th>Statistical</th>
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<tbody>
<tr>
<td></td>
<td>C</td>
<td>F</td>
<td>W</td>
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<tr>
<td><strong>Total $^{13}$C-incorporation into PLFA</strong> (µg $^{13}$C per mg added $^{13}$C g$^{-1}$ soil)</td>
<td></td>
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<tr>
<td>Glucose</td>
<td>1.48±0.13</td>
<td>1.35±0.06</td>
<td>1.19±0.12</td>
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<tr>
<td>Glycine</td>
<td>0.92±0.11</td>
<td>0.85±0.06</td>
<td>0.85±0.12</td>
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<tr>
<td>Acetic acid</td>
<td>2.38±0.24</td>
<td>2.50±0.11</td>
<td>2.20±0.08</td>
</tr>
<tr>
<td>Starch</td>
<td>0.40±0.06</td>
<td>0.33±0.06</td>
<td>0.28±0.03</td>
</tr>
<tr>
<td>Vanillin</td>
<td>0.32±0.03</td>
<td>0.41±0.01</td>
<td>0.24±0.01</td>
</tr>
<tr>
<td><strong>Total $^{13}$C-incorporation into NLFA</strong> (µg $^{13}$C per mg added $^{13}$C g$^{-1}$ soil)</td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>Glucose</td>
<td>0.51±0.06</td>
<td>0.70±0.08</td>
<td>0.51±0.04</td>
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<tr>
<td>Glycine</td>
<td>0.15±0.02</td>
<td>0.14±0.02</td>
<td>0.23±0.05</td>
</tr>
<tr>
<td>Acetic acid</td>
<td>1.27±0.11</td>
<td>1.45±0.13</td>
<td>1.51±0.12</td>
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<td>Starch</td>
<td>0.06±0.01</td>
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<tr>
<td>Vanillin</td>
<td>0.05±0.01</td>
<td>0.12±0.01</td>
<td>0.10±0.04</td>
</tr>
<tr>
<td><strong>NLFA/PLFA for $^{13}$C-incorporation from</strong></td>
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<tr>
<td>Glucose to 18:2ω6.9</td>
<td>0.09±0.01</td>
<td>0.15±0.03</td>
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</tr>
<tr>
<td>Glucose to 18:1ω9</td>
<td>1.09±0.16</td>
<td>1.78±0.15</td>
<td>1.65±0.25</td>
</tr>
<tr>
<td>Acetic acid to 18:2ω6.9</td>
<td>0.16±0.01</td>
<td>0.12±0.02</td>
<td>0.12±0.02</td>
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<tr>
<td>Acetic acid to 18:1ω9</td>
<td>1.65±0.23</td>
<td>1.53±0.20</td>
<td>2.21±0.42</td>
</tr>
</tbody>
</table>

*C*: control; F: fertilization; W: warming; W+F: warming+fertilization.

n.s.: no significant effects at P<0.01.

Significant effects of factors warming (W), fertilizer addition (F), time (T) and their interactions are shown at **P<0.01, ***P<0.001.

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Figure 4. Principal component analysis (PCA) of the relative $^{13}$C-incorporation of added $^{13}$C-labelled glucose, glycine, acetic acid, starch and vanillin into phospholipid fatty acids (PLFAs) of tundra heath soil. Scores (mean ± SE) of the principal components (PC) 1 and 2 showing effects of warming and fertilization, and differences between the incubation times (1 day, small symbols drawn with black line; 7 days, large symbols drawn with grey line). The PCAs were run separately for each substrate. The explained variances were 33–46% for PC1 and 18–27% for PC2. For glucose, glycine, acetic acid and vanillin P<0.001 for Fertilization and Time effect on the PC scores (Linear Mixed Models with Warming, Fertilization and Time as fixed factors and Block as a random factor). For starch P<0.01 for Warming, P<0.001 for Fertilization and Time, and P<0.01 for Warming×Fertilization.

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Discussion

Fertilization of the subarctic heath for 18 years had increased the total PLFA concentration in soil by 31%, which indicates increased microbial biomass. This increase is in agreement with the results after 15 treatment years, when fertilization also had increased the total PLFA concentration, however only by 16% [5]. In contrast to the reduction of total PLFA by the warming treatment after 15 years [5], warming for the additional three years up to the sampling for the present study had led to a significantly increased total PLFA concentration. This implies that the total microbial biomass responses in the warming treatment now, for the first time, resemble those in the fertilization treatment [21]. The result probably reflects the long-term increase in the total cover of vascular vegetation. After 22 years, both warming and fertilization treatments had 35% higher vascular plant cover than the control [16], in contrast to previous measurements after 5 and 10 years, with varying responses in the two treatments [22]. Increased primary production typically leads to an increase in soil microbial biomass [23], thanks to enhanced substrate inputs.

Based on a PCA of the PLFA profiles, the microbial community composition in the fertilized soil was characterized by a significantly higher abundance of the fungal biomarker, 18:2ω6,9, as compared to the unfertilized soil, which was similar to the results after 15 treatment years [5]. The fertilization-induced increase in fungi was further evidenced by the significantly higher total concentration of NLFAs, which are mainly found in eukaryotic organisms and thus are indicative of
fungi in soil [24]. More specifically, the fungal marker NLFA, 18:2\text{6,9} and 18:1\text{6,9} [24] had a significantly higher relative abundance in the fertilized than in the unfertilized soil. Rimman et al. [5] discussed possible effects of adding N on soil fungi, also comparing with other habitats like confierous forest soil. They suggested that the fertilizer-induced increase in fungi in this peaty subarctic soil resulted from 1) an alleviated N or P limitation of the fungal exoenzyme production, 2) a greater predation pressure on the bacteria than on the fungal community or 3) a higher abundance of mycorrhizal fungal mycelium. An altered community composition of ectomycorrhizal fungi was also found in a fertilization experiment in Arctic soils in North America [25].

Direct effects of a higher temperature on the PLFA pattern is well studied, with increased length and lower degree of unsaturation of fatty acids at higher temperatures [26]. Such changes were not found in the present study, indicating no direct effect of temperature. Instead warming caused only a minimal shift in the fatty acid profiles, and the effect of warming overrode that of fertilization as the combined warming-fertilization treatment had a similar PLFA profile to warming alone. Thus, although warming had increased the soil microbial biomass (total PLFA) indicating similar effects as fertilization, the different responses in the community composition suggest that the mechanisms behind the increase in microbial biomass under these treatments are different. This is in line with the vegetation responses, which show that despite the identical increase in the total cover of vascular plants in warming and fertilization treatments, the responses in the dominant plant species under these treatment were contrasting. While warming enhanced the cover of ectomycorrhizal *Betula nana*, fertilization benefited the ericoid mycorrhizal *Empetrum hermaphroditum* [16]. It is highly unlikely that the air warming by open-top chambers has had any direct effect on soil microbial communities in the long term, because the expansion of the vascular plant cover has over time led to shading and no soil temperature differences between the control and warming treatments [27].

The most obvious difference between the warming and fertilization treatments is that the supplied NPK fertilization provides the microbes with extra N and P, alleviating limitation by these nutrients, whereas soil warming increases mineralization of carbon as well as other nutrients. Earlier results have shown that microbes are N-limited or limited by both C and N in northern soils [6,28], including soils from the Abisko region [3,5]. Findings from another experiment in Abisko suggest, however, that warming shifts the originally N-limited soil microbial community slightly towards carbon limitation, although no significant changes in mineralization are observed [3]. This indicates that the warming-induced increase in net mineralization may be small, as also suggested by in situ buried bag studies at the experimental site [29,30].

Utilization of the amended substrates by soil microbial communities was mainly affected by fertilization, while warming only had few effects. The most pronounced response was that the fertilized soil was characterized by a relatively higher 13C-incorporation into fungi than into bacteria as compared with the unfertilized soil, irrespective of the added substrate. Incorporation into the fungal NLFA 18:2\text{6,9} followed a similar response, but because of high variance, only vanillin showed a statistically significant fertilization effect. The 13C-recovery in the fatty acid 18:1\text{6,9}, which also indicates fungi [24], was increased by fertilization for vanillin (both PLFA and NLFA) and starch (only NLFA).

Fungi store carbon in excess as triacylglycerols, which are detected in the NLFA fraction [24]. Allocation to fungal energy storage can thus be estimated with NLFA-to-PLFA ratios, and a lower ratio after adding C-rich, labile, substrates indicates available N in the soil [24]. Contrary to our hypothesis, there were no significant fertilization effects on the NLFA-to-PLFA ratios for 13C-incorporation. However, the NLFA-to-PLFA ratios for 13C-incorporation from acetic acid into 18:2\text{6,9} and 18:1\text{6,9} were higher in the warmed soil but unaffected by fertilization compared to the control. The increased allocation to storage under warming suggests that the increased temperature actually led to limitation of growth by the lack of nutrients when a simple carbon-rich substrate, acetic acid, was easily available. The contrasting responses of warming and fertilization clearly show that fertilization cannot be used as a proxy for effects of warming, as also indicated by the PLFA pattern.

That fertilizer addition increased 13C-incorporation from vanillin (a model substance indicating lignin depolymerisation), while warming had no significant effect, is additional evidence for this conclusion. This is also in agreement with the results of Neff et al. [31], who observed that fertilization of an alpine meadow had triggered increased decomposition of plant-derived lignin compounds. White-rot fungi of coniferous forest soil are known to down-regulate the production of lignolytic extracellular enzymes under high soil N concentrations [32,33], suggesting that N fertilization should result in decreased lignin degradation. This was not the case in the present study. However, the increased vanillin utilization was mainly due to the stimulated uptake of 13C to the cyclic and fungal fatty acids, suggesting that both bacteria and fungi were involved in vanillin degradation in the soil studied.

The pattern of 13C incorporation from starch into PLFAs differed from the other substrates, which was also reported earlier [34]. Thus, although starch degradation is fairly common among microorganisms [35], a different subset of microorganisms appears to degrade starch as compared to the other studied substrates. Also, in contrast to the other substrates used in this study, utilization of starch was significantly affected by both fertilization and warming. More specifically, the incorporation of 13C from starch into indicators of Gram-positive bacteria (16:0) and actinomycetes (10Me16:0 for fertilization and 10Me18:0 for warming) increased by the treatments. It appears that the starch-degrading microorganisms were favored by the changes that had occurred both in the fertilized and warmed plots, possibly by the increase in deciduous shrubs and the concomitant increase in leaf litter [8,15].

Increased biomass and activity of soil fungi relative to those of bacteria in response to enhanced nutrient availability may lead to altered ecosystem functioning. Most importantly, an increased importance of fungi relative to bacteria has been suggested to result in an increased carbon sequestration in soil [36], but see also discussion by [37]. Because of the enormous size of the arctic soil carbon pool, even tiny changes in soil carbon processing have a potential to feed back on climate change. Effects of climate change on microbial use of carbon substrates would influence soil carbon storage both directly, as altered efflux of carbon from the soil, and indirectly, via priming effects (for a review, see [30]) that could lead to altered degradation of older carbon.

Several of our results indicate that – at least in the 18-year time horizon – warming does not result in similar effects as experimentally increased soil nutrient availability. Thus, climate warming will not increase carbon sequestration in tundra heaths via beneficial effects on soil fungi to the same degree as fertilization unless soil nutrient availability increases more than the increase in moderately warmer soil. As the N deposition rates in the region north of 60° N are low [39], a pronounced increase in nutrient availability via atmospheric deposition is, however, unlikely.
Materials and Methods

Experimental Design and Soil Sampling

A long-term climate change simulation experiment has been maintained on a tree-line heath at Abisko, northern Sweden (68°21′N, 18°49′E, 450 m a.s.l.), since 1989. The vegetation is dominated by ericoid dwarf shrubs (mainly Cassiope tetragona (L.) D. Don) and mosses, with deciduous shrubs and graminoids as subdominants. The experiment host various treatments (originally described in [21]) of which 1) warming, 2) fertilization, 3) the combination of warming and fertilization and 4) unmanipulated control plots were selected for the present study. For the warming treatment, 3–4°C air and 1–2°C soil temperature (4 cm depth; [40,41]) increases were achieved with the help of dome-shaped open-top plastic greenhouses, that were erected on the plots for the growing season from early June to late August. The fertilization treatment simulated enhanced nutrient availability and it was maintained by annual additions of 10 g m⁻² N, 2.6 g m⁻² P and 9 g m⁻² K as NH₄NO₃, KH₂PO₄ and KCl (except 1989, when half of this amount was applied, and 1993 and 1998, when no fertilization was performed). Each of the plots covered a 1.2 x 1.2 m area and was randomly distributed in six blocks, i.e. there were six replicate plots of each treatment. No specific permits were required for the described field studies, as the location is not privately-owned and not protected and the field studies did not involve endangered or protected species.

Three soil cores were taken from each field plot on August 24, 2006, i.e. after 18 years of treatments. The 5-cm-deep cores with a diameter of 4 cm were mixed into a single sample per treatment plot (total volume of 65 cm³), while removing roots and stones by hand sorting. The soil has a c. 15 cm deep organic layer with a pH of 7.1, which has not been affected by the treatments [41] and a soil organic matter content of 89% in the unfertilized and 92% in the fertilized soil in the top 5 cm [5]. The organic layer rests on rocky mineral soil without permafrost.

Incubation with ¹³C-labelled Substrates

Uptake of label from ¹³C-labelled substrates into fatty acids was determined by SIP-PLFA as described by [34]. Five different substrates were used: universally ¹³C-labelled glucose (99%), glycine (99%) and starch (98%), ring-labelled vanillin (99%) and acetic acid, starch and vanillin into the PLFA 18:2 v9,9 is mainly of fungal origin (see discussion in [47]), and the NLFA 18:1 v7 is of fungal origin, and often found to increase under excess of carbon [24].

Statistical Analysis

The fatty acid concentrations and the total ¹³C-incorporation to fatty acids were analyzed for treatment effects by Linear Mixed Models of SPSS 14.0 for Windows with warming, fertilization, substrate and incubation time as fixed factors. The block variable from the field experiment was used as a random factor. As the differences between the ¹³C-labelled substrates were not of primary interest in the present study and have been described before [34], the analysis was run separately for each substrate. To reduce the chance for type I errors, the alpha-level of 0.01 was used to indicate statistical significance.

The unit-variance scaled and centered fatty acid mole percentage profiles and the patterns of ¹³C-incorporation were subjected to principal component analyses (PCA) using Simca-P 11.0 (Umetrics, Umeå, Sweden). The extracted principal components (PC) were analyzed for treatment effects using Linear Mixed Models in a similar manner as described above.

Supporting Information

Figure S1 Incorporation of ¹³C from glucose, glycine, acetic acid, starch and vanillin into the PLFA 18:206:9. Tundra heath soils from a field experiment with long-term warming and fertilization treatments were incubated 1 and 7 days. The bars represent mean±SE. Note different y-axis scales.

Figure S2 Incorporation of ¹³C from glucose, glycine, acetic acid, starch and vanillin into the PLFA 18:109. Tundra heath soils from a field experiment with long-term warming and fertilization treatments were incubated 1 and 7 days. The bars represent mean±SE. Note different y-axis scales.
Significant effects of factors warming (W), fertilizer addition (F), time (T) and their interactions for each substrate are shown at **P < 0.01, ***P < 0.001 (Linear Mixed Models with Warming, Fertilization and Time as fixed factors and Block as a random factor).

(TIF)

Figure S3 Incorporation of $^{13}$C from glucose, glycine, acetic acid, starch and vanillin into the PLFA cy19:0. Tundra heath soils from a field experiment with long-term warming and fertilization treatments were incubated 1 and 7 days. The bars represent mean±SE. Note different y-axis scales. Significant effects of factors warming (W), fertilizer addition (F), time (T) and their interactions for each substrate are shown at **P < 0.01, ***P < 0.001 (Linear Mixed Models with Warming, Fertilization and Time as fixed factors and Block as a random factor).

(TIF)

Figure S4 Incorporation of $^{13}$C from glucose, glycine, acetic acid, starch and vanillin into the PLFA 18:1o7. Tundra heath soils from a field experiment with long-term warming and fertilization treatments were incubated 1 and 7 days. The bars represent mean±SE. Note different y-axis scales. Significant effects of factors warming (W), fertilizer addition (F), time (T) and their interactions for each substrate are shown at **P < 0.01, ***P < 0.001 (Linear Mixed Models with Warming, Fertilization and Time as fixed factors and Block as a random factor).

(TIF)

Figure S5 Incorporation of $^{13}$C from glucose, glycine, acetic acid, starch and vanillin into the PLFA 15:0. Tundra heath soils from a field experiment with long-term warming and fertilization treatments were incubated 1 and 7 days. The bars represent mean±SE. Note different y-axis scales. Significant effects of factors warming (W), fertilizer addition (F), time (T) and their interactions for each substrate are shown at **P < 0.01, ***P < 0.001 (Linear Mixed Models with Warming, Fertilization and Time as fixed factors and Block as a random factor).

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Figure S6 Incorporation of $^{13}$C from glucose, glycine, acetic acid, starch and vanillin into the PLFA 16:0. Tundra heath soils from a field experiment with long-term warming and fertilization treatments were incubated 1 and 7 days. The bars represent mean±SE. Note different y-axis scales. Significant effects of factors warming (W), fertilizer addition (F), time (T) and their interactions for each substrate are shown at **P < 0.01, ***P < 0.001 (Linear Mixed Models with Warming, Fertilization and Time as fixed factors and Block as a random factor).

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Author Contributions

Conceived and designed the experiments: RR AM EB. Performed the experiments: RR. Analyzed the data: RR EB. Wrote the paper: RR AM EB.

References


