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Abrupt permafrost thaw triggers activity of copiotrophs and microbiome predators

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Abstract

Permafrost soils store a substantial part of the global soil carbon and nitrogen. However, global warming causes abrupt erosion and gradual thaw, which make these stocks vulnerable to microbial decomposition into greenhouse gases. Here, we investigated the microbial response to abrupt thaw in situ permafrost thaw. We sequenced the total RNA of a 1 m deep soil core consisting of up to 2500-year-old permafrost material from an active abrupt erosion site. We analysed the microbial community in the active layer soil, the recently thawed, and the intact permafrost, and found maximum RNA:DNA ratios in recently thawed permafrost indicating a high microbial activity. In thawed permafrost, potentially copiotrophic Burkholderiales and Sphingobacteriales, but also microbiome predators dominated the community. Overall, both thaw-dependent and long-term soil properties significantly correlated with changes in community composition, as did microbiome predator abundance. Bacterial predators were dominated in shallower depths by Myxococcales, while protozoa, especially Cercozoa and Ciliophora, almost tripled in relative abundance in thawed layers. Our findings highlight the ecological importance of a diverse interkingdom and active microbial community highly abundant in abruptly thawing permafrost, as well as predation as potential biological control mechanisms.

Keywords: abrupt erosion; copiotrophic; permafrost; protozoa; transcriptomics

Introduction

Permafrost soils are among the most vulnerable ecosystems of the globe, both in terms of biodiversity and potential degradation in response to global warming (Abbott et al. 2022, IPCC 2021). Permafrost is found in historically cold regions and remains frozen for at least 2 consecutive years. It consists of a seasonally thawed active layer and permafrost below, often remaining frozen for millennia. These conditions selected for a unique but still elusive microbial communities under constant freezing conditions (Jansson and Tas 2014). Intact permafrost—due to slow microbial at long-term stable freezing temperatures—stores 1100–1500 Pg C, over half of all global soil carbon (C) (Tarnocai et al. 2009, Hugelius et al. 2014), as highly recalcitrant organic matter (Farmentier et al. 2017). As ambient temperatures in the Arctic increase up to four times faster than the global average (Rantanen et al. 2022), active layers gradually thaw deeper into former permafrost (AMAP 2021). In the long term, this enables a deeper rooting vegetation and potential microbial remineralisation of permafrost carbon (Schuur and Mack 2018). Additionally, warming causes ground ice to melt and soils to collapse abruptly, modelled to affect up to half of all permafrost carbon by 2100 (Turetsky et al. 2020).

In ancient microaggregates and brine channels, permafrost harbours resistant taxa that exhibit slow growth rates, higher metabolic versatility, and an ability to perform syntrophy (Allison and Martiny 2008, Shade et al. 2012), often functionally constrained due to environmental limitations and slow reproduction rates (Monteux et al. 2020). The overlaying thawed soils selected for highly resilient taxa with high growth rates, as temperature seasonally fluctuates (Allison and Martiny 2008, Bardgett and Caruso 2020). Upon thaw, permafrost soil communities, then can become functionally redundant (Nannipieri et al. 2017), both during in situ (Monteux et al. 2018) and experimental thaw (Monteux et al. 2020). Thaw-induced perturbation of the ground can cause migrations of microbial taxa from the original active layer towards deeper soils, called coalescence. This can be induced through rapid population growth of copiotrophic taxa and can increase carbon and nitrogen (N) release from thawed permafrost soil (Monteux et al. 2020). These microbial responses to thaw could stem from the revival of former resting stages, such as cysts and endospores (Lennon and Jones 2011), which upon suitable thermal and moisture conditions can lead to locally increased microbial activity.

While so far not described in thawing permafrost soils, the complexity of trophic relations plays a key role in temperate soil carbon mineralisation (Bardgett and van der Putten 2014). Eukaryotes and their importance in permafrost are particularly under-
studied due to biases and difficulties related to the use of gene-specific primers (Harder et al. 2016), but the use of total RNA sequencing enables a less biased understanding of the putatively active total community. Particularly, the predation on bacteria can act as control and driver of the soil microbiome (Thakur and Geisen 2019, Geisen et al. 2021). Organisms feeding on bacteria, called bacterivores, include predatory eukaryotes, such as protozoa (sensu single cell predatory protists; Geisen et al. 2018), nematodes, rotifers, and tardigrades (Coleman and Wall 2015), as well as bacterial taxa, such as the myxobacterial phyla Myxococcota and Bdellovirionibota. Moreover, the selective removal of bacterial cells by predators controls both the bacterial turnover and community composition (Trap et al. 2016). Lysis of prey cells and incomplete mineralisation of organic C and N by protozoa enhance the recycling and distribution of nutrients (Bonkowski 2004). Although microbial predator presence, such as Myxococcota, was documented before in permafrost (Maliard and Pearce 2018, Schostag et al. 2019, Scheel et al. 2022), their relevance in Arctic soils is still elusive.

In this study, we investigated the whole active microbial community of recently thawed ancient permafrost and tested how strongly biotic or abiotic factors impact the microbial community in different stages of thaw. Due to their shorter doubling time, we furthermore hypothesised that the freshly thawed soils support higher abundances of copiotrophic taxa. We then discuss their potential origin from the active layer or reactivated permafrost resting stages as well as how their biomass could support a potential bacterivore cascade. These insights can help us understand the spatiotemporal and trophic dynamics of permafrost microbial ecology and estimate the ecological response of gradual and rapid erosion as a potential key driver of both permafrost carbon vulnerability.

Materials and methods

Soil sampling

Sampling took place in 2020 in Zackenberg valley, NE Greenland (74°30’N 20°30’W, Fig 1A–B). This wide lowland valley is dominated by continuous permafrost and a vegetation of wet hummocky fens, low shrub, and graminoids (Elberling et al. 2008). Average temperatures varied between −2°C in summer and −14°C in winter between 1997 and 2006 (Christiansen et al. 2008) The active layer seasonally thaws between 40 cm and 2 m deep, but an increase of 0.77 cm per year when including data from 1995 to 2020 (Westermann et al. 2015, Westergaard-Nielsen et al. 2018).

The permafrost soil surface collapsed abruptly in 2018 as a formerly described recent, thermal erosion gully developed in the vicinity of the Zackenberg Research Station (Christensen et al. 2020, Scheller et al. 2021, Scheel et al. 2022). Below the active layer (AL) at 40 cm depth, an ice lens had melted in 2019, creating a first transition zone depth until 70 cm depth (TZ1). In 2020, thaw reached until 90 cm depth in 2020 (TZ2), while below 90 cm depth intact permafrost (PF) persisted (Fig 1C–D). In 2020, three replicate soil samples were taken aecpially per 10 cm intervals until a depth of 1 m, resulting in 30 samples. Due to different RNA stability at varying freezing temperatures (Schostag et al. 2020) and laboratory limitations in the sampling station, the samples were stored at −20°C until transported frozen to Denmark, where they were stored at −80°C.

Physicochemical soil analysis

Physical soil properties were determined in technical triplicate as described in Scheel et al. (2022) from thawed 10-cm soil horizon samples until 100 cm depth. First after air-drying at 70°C for 48 h, followed by burning at 450°C for 2 h, the samples were weighed to determine the loss of relative weight-based soil water (H2O) and organic carbon (OM, Wilke 2005). The pH was measured after adding 50 ml of 1 M KCl to 10 ml of air-dried soil samples with a Mettler Toledo FiveEasy Plus™ pH Meter (Mettler Toledo GmbH, Giemsen, Germany).

Radiocarbon dating was performed per 10-cm horizon by sifting thawed soil with a 0.5-mm sieving retaining macro plant residues and excluding roots. Per depth, triplicates were pooled, treated with HCl and NaOH, graphitised, and 14C isotope activity was measured using an accelerator mass spectrometer (Radiocarbon Dating Laboratory, Lund University, Lund, Sweden). The resulting age was calibrated with IntCal13 (Reimer et al. 2016) to 14C years in BP (before present = AD 1950) and the Levin post-Bomb calibration (Levin and Kromer 2016) for results in FM (fraction modern) after 1963.

Nucleic acid co-extraction, library preparation, and sequencing

Deep-frozen samples were homogenised in antiseptic mortars. Then we co-extracted the total RNA and DNA of the biological replicates on up to 0.35 g frozen soil sample with the NucleoBond RNA Soil Mini kit (Macherey-Nagel GmbH & Co KG Dueren, Germany) according to the manufacturer’s protocol. We used G2 DNA/RNA Enhancer infused 1, 4 mm beads (Ampliqon, Odense, Denmark) instead of the ones provided with the kit. To remove potential DNA, the RNA extracts were treated with the DNase Max Kit (QIAGEN), following the manufacturer’s protocol. Both the removal of DNA and final RNA concentrations were evaluated and confirmed using a Qubit® 4 Fluorometer (Thermo Fisher Scientific, Life Technologies, Roskilde, Denmark). We quality-assessed the final extracts with a TapeStation 4150 (Agilent Technologies, Santa Clara, CA, USA) using a high-sensitivity assay. Resulting RNA integrity numbers (RIN) were low (1.0 ± 0.6, Supplementary Table 1) and samples from 80 to 100 cm depth had RNA concentrations below detection limit. Nucleic acid concentrations were normalised for sample weight and used for extraction ratios of extracted RNA to DNA (RNA:DNA, Table 1). We fragmented the RNA, synthesised cDNA, and processed the cDNA with the NEBNext Ultra II Directional RNA Library Prep Kit and the NEBNext Multiplex Oligos for Illumina (New England BioLabs, Ipswich, MA, USA), following the manufacturer’s protocol. The resulting samples were pooled into equimolar metatranscriptome libraries to secure even sequencing coverage. We performed the sequencing in-house (Department of Environmental Science, Aarhus University, Denmark) on an Illumina NextSeq 500 with a v2.5 high-throughput 300 cycles kit (both Illumina, San Diego, CA, USA).

Bioinformatic processing

The 352 Mio. raw paired-end Illumina reads (SRA accession number: PRJNA939404, Table 1) were quality-controlled with TrimGalore (https://www.bioinformatics.babraham.ac.uk/projects/trim_galore/) and filtered by removing adapters and short reads (<60 nt). These sequences were then classified (sorted) into small subunit (SSU) rRNA, large subunit (LSU) rRNA, and non-rRNA sequences by aligning to SILVA 138.1 SSU Ref NR 99 and SILVA 138.1 LSU Ref NR 99 using SortMeRNA (Kopylova et al. 2012). The SSU reads were assembled into full-length SSU rRNA contigs with MetaRib (Xue et al. 2020). These rRNA contigs were taxonomically classified as described by Anwar and colleagues (Anwar et al. 2019), using CREST4 (Lanzen et al. 2012) against the
SILVA database v.138 (Quast et al. 2013) and PR2 database (Guillou et al. 2013). The resulting annotation was partly outdated and corrected per contig (Supplementary Methods). The rRNA reads were mapped to resulting EMIRGE contigs using BWA (Li and Durbin 2009), resulting in taxonomically annotated rRNA contig abundance across the 30 samples. The automatic pipeline used is available (Campuzano Jiménez 2023).

### rRNA processing

The taxonomic sequence abundance was scanned for previously documented protozoan taxa: Amoebozoa, SAR, Euglenozoa, Foraminifera, and Heterolobosa (Geisen et al. 2015) and Endomyxa, Telonemia, Malawimonadidae, and Choanoflagellida. Bacterial predators included Myxococcota (excluding the genus

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**Figure 1.** (A) Sampling site Zackenberg in Northeast Greenland, credit: Google Earth. (B) Sampling site in 2018 after initial permafrost collapse. (C) Soil profile from the surface until still frozen depth at 90 cm during sampling in 2020. (D) Scheme of abrupt permafrost thaw (indicated with blue arrows), depicting the soil profile until the permafrost (PF) layer at 90–100 cm depth from the moment of collapse in 2018 with visible ice lens below long-term active layer (AL) to the formation of transition zones (TZ = thawed permafrost).

**Table 1.** Soil properties include $^{14}$C (radiocarbon) dating results (* = IM, **BP), relative weight-based soil moisture (H$_2$O), relative weight-based soil organic matter content (SOM), pH and layer, as defined by thawing processes as active layer (AL); deepest thaw in 2019: transition zone 1 (TZ1) and 2020: TZ2, and permafrost (PF).

<table>
<thead>
<tr>
<th>Depth [cm]</th>
<th>$^{14}$C [fM **BP]</th>
<th>H$_2$O [%]</th>
<th>SOM [%]</th>
<th>pH</th>
<th>Layer</th>
<th>DNA ng/gDW</th>
<th>RNA ng/gDW</th>
<th>RNA:DNA</th>
<th>Total reads</th>
</tr>
</thead>
<tbody>
<tr>
<td>0–10</td>
<td>1.04*</td>
<td>28.80 ± 3.38</td>
<td>8.73 ± 1.82</td>
<td>4.22 ± 0.03</td>
<td>AL</td>
<td>835.55 ± 391.99</td>
<td>357.46 ± 171.55</td>
<td>0.46 ± 0.16</td>
<td>1.25E+07</td>
</tr>
<tr>
<td>10–20</td>
<td>1.13*</td>
<td>22.52 ± 0.73</td>
<td>5.39 ± 0.83</td>
<td>4.02 ± 0.05</td>
<td>AL</td>
<td>826.65 ± 93.13</td>
<td>72.99 ± 8.38</td>
<td>1.78 ± 1.28</td>
<td>4.85E+07</td>
</tr>
<tr>
<td>20–30</td>
<td>1.16*</td>
<td>22.05 ± 3.30</td>
<td>10.19 ± 3.22</td>
<td>4.29 ± 0.01</td>
<td>AL</td>
<td>561.38 ± 571.54</td>
<td>91.19 ± 1.99</td>
<td>0.29 ± 0.19</td>
<td>2.25E+07</td>
</tr>
<tr>
<td>30–40</td>
<td>1.2*</td>
<td>26.57 ± 0.83</td>
<td>13.68 ± 0.44</td>
<td>4.25 ± 0.01</td>
<td>AL</td>
<td>186.67 ± 33.49</td>
<td>86.43 ± 23.06</td>
<td>0.49 ± 0.20</td>
<td>6.0E+07</td>
</tr>
<tr>
<td>40–50</td>
<td>2635**</td>
<td>7.73 ± 1.58</td>
<td>2.59 ± 0.68</td>
<td>4.63 ± 0.03</td>
<td>TZ1</td>
<td>54.68 ± 60.24</td>
<td>40.66 ± 42.59</td>
<td>0.85 ± 0.91</td>
<td>2.39E+07</td>
</tr>
<tr>
<td>50–60</td>
<td>3770**</td>
<td>15.61 ± 6.28</td>
<td>2.93 ± 0.30</td>
<td>4.13 ± 0.02</td>
<td>TZ1</td>
<td>10.99 ± 9.45</td>
<td>40.51 ± 10.58</td>
<td>0.93 ± 11.03</td>
<td>7.22E+07</td>
</tr>
<tr>
<td>60–70</td>
<td>26500**</td>
<td>8.72 ± 2.53</td>
<td>1.52 ± 0.16</td>
<td>4.86 ± 0.05</td>
<td>TZ2</td>
<td>8.66 ± 11.41</td>
<td>36.19 ± 4.39</td>
<td>11.98 ± 9.47</td>
<td>4.90E+07</td>
</tr>
<tr>
<td>70–80</td>
<td>22100**</td>
<td>6.35 ± 0.60</td>
<td>1.10 ± 0.24</td>
<td>4.48 ± 0.01</td>
<td>TZ2</td>
<td>2.11 ± 11.97</td>
<td>17.94 ± 31.07</td>
<td>18.48 ± 32.02</td>
<td>1.49E+07</td>
</tr>
<tr>
<td>80–90</td>
<td>26200**</td>
<td>7.96 ± 0.61</td>
<td>1.00 ± 0.20</td>
<td>4.57 ± 0.01</td>
<td>PF</td>
<td>1.04 ± 1.80</td>
<td>NA</td>
<td>NA</td>
<td>3.46E+07</td>
</tr>
<tr>
<td>90–100</td>
<td>26200**</td>
<td>7.80 ± 0.59</td>
<td>1.04 ± 0.26</td>
<td>4.91 ± 0.01</td>
<td>PF</td>
<td>0.96 ± 1.66</td>
<td>NA</td>
<td>NA</td>
<td>1.40E+07</td>
</tr>
</tbody>
</table>

Extraction and sequencing output included the calculated RNA:DNA ratio based on co-extraction of nucleic acids in ng yield per gram dry weight (gDW). The cumulative sum of clean total RNA sequencing reads per depth. Standard deviations for triplicates per depth are indicated where available (a).
Sorangium), Bdellovibrio nata, Lysobacter, Daptobacter, and Vampiro- 
coccus (Petters et al. 2021, adapted according to SILVA v.138 taxon- 
omy). The rRNA relative abundance was calculated and collapsed taxonomically into prokaryotes, eukaryotes, and further into bacte-
rial predators and protozoa.

**Statistical analysis and data processing**

For statistical analysis, macroeukaryotes were removed from the dataset, to only depict microbial interactions. We performed com-
munity analysis with the R software v.4.2.2 in R studio (RSu-
dio Team 2020, R Core Team 2022), using the phyloseq v. 1.42.0 (McMurdie and Holmes 2013) and vegan v. 2.6.4 packages (Oksa-
en et al. 2022). We excluded the triplicate 1 of 30 (for the depth 0–10 cm) for all downstream analysis, due to significantly low number of reads (360 opposed to an average 12 Mio reads per triplicate). Shannon (S) alpha diversity was calculated on the total number of rRNA contigs and abundance of sequence reads mapped per sample (Supplementary Table 1). The significance of environmental data (age, pH, SOM, H₂O, and layer) and of bacte-
rial and protozoan predator abundance were tested with the anova function (PERMANOVA, 999 permutations) on Bray–Curtis dissimilar-
ities of relative abundance per taxon within the sample to ac-
count for different read coverage across samples. For predation ef-
effects, we performed the tests on Bray–Curtis dissimilarities of only nonpredatory prokaryotic and eukaryotic communities. We per-
formed consecutive Tukey’s HSD post-hoc tests to determine sig-
nificant difference between taxonomic groups with layer and age as explanatory variables. Variance partitioning was performed on PERMANOVA results for thaw-related (layer and H₂O), long-term soil (SOM, pH, age) as well as biotic (bacterial and protozoan pre-
dation) parameters. Nonmetric multidimensional scaling (NMDS) plots were ordinated graphically using Bray–Curtis dissimilarities between samples of the subset communities. Potential predator–
prey interactions were investigated using the Sparse Inverse Co-
variance estimation for Ecological Association and Statistical In-
ference (SPIEC-EASI) analysis (Kurtz et al. 2015) as can be found in the Supplementary Materials.

**Results**

**Physicochemical soil properties**

Radiocarbon results differed greatly with depth, indicating three age categories, that were used for downstream statistical analy-

sis. A more recent upper consisted of organic material and silt (AY), a deeper 2635–3770-year-old inorganic silt layer (AM), and the underlying deepest 22100–26500-year-old layer of inorganic sand and gravel (AO; Table 1). Soil moisture was highest in the active layer until 40 cm depth and then stayed rather stable at 6.35%–8.72%, while pH stayed stable throughout between 4.02 and 4.91 (Table 1). Despite low biomass (174.4 ± 334.7 ng/dw 
DNA; 47.3 ± 111.7 ng/dw RNA), extraction of RNA and DNA was successful in all triplicates and RNA:DNA ratios indicated higher values in the transition zones (Supplementary Fig. 1; Supplementary Table 1).

**Total rRNA community composition**

In total, 59899 full-length rRNA gene contigs were constructed with an average length of 1474 bp. On average, 74.6% of all reads could be annotated as SSU rRNA and 65.5% of the trimmed reads mapped back to the assembled full-length rRNA genes. Of these, 59899 were annotated on domain level, but annotation suc-
cess decreased above family level (3353 contigs = 56.2% anno-
tated, Supplementary Table 1). Overall, 5582 rRNA contigs were assigned as Bacteria (79.6% ± 0.1% of total counts), 374 as Eu-
karya (14.8% ± 0.2%), and 16 as Archaea (0.1% ± 0.0%). Om-
nipresent in each sample were 3900 rRNA contigs (65.1% of con-
tigs, 91.7% of all counts). Yet, 2635 rRNA contigs were most abund-
ant in active layer (44.0% of contigs, 26.9% of all counts). Only 573 and 294 rRNA contigs were most abundant specifically in TZ1 and T22, compromising only 9.6% and 4.9% of contigs, but 12.2% and 9.6% of all counts, respectively. Further, 245 rRNA contigs were most abundant in permafrost (4.1% of contigs, 6.2% of all counts).

**Prokaryotic community composition and distribution with depth**

The prokaryotic community was dominated by Proteobacteria (21.8% ± 0.1% of total counts), Actinobacteriota (15.0% ± 0.1% of total counts), Acidobacteriota (8.1% ± 0.0% of total counts), Bac-
teroidota (6.3% ± 0.1% of total counts), and Myxococciota (6.0% ± 0.1% of total counts, Fig 3A). Further, S1 bacterial phyla (including 16 Candidata phyla) and 7 archaenal phyla were present. Proteobac-
teria increased in relative abundance between 50 and 90 cm, the depth of the two thaw layers. This phylum was dominated by the gamma-proteobacterial Burkholideriales (14% ± 0.1% of total counts) with maximum abundances at 70 and 90 cm depth, respectively. The abundance of Actinobacteriota decrease with depth, but peaked in permafrost, while Acidobacteriota where most abundant in the active layer. Bacteroidota abundances in-
creased with depth, especially within 70–100 cm, responding to the layers T22 and permafrost. Especially, Sphingobacterial rRNA contigs constituted 3.0% ± 0.2% of all counts and reached maximum abundances at 70–80 cm.

**Microeukaryotic permafrost community**

Eukaryotes were dominated by Plantae rRNA contigs (7.5% ± 0.6% of total counts). Together with Metazoa (0.5% ± 0.3%), such as Arthropoda and Tardigrada rRNA contigs, macroeukaryotes were removed from further statistical analysis. Fungi were over-
all present at low relative abundance of 0.6% ± 0.0% of the total.
Figure 3. All figures are given for subsets of the total community with prokaryotes, bacterial predators, eukaryotes, and protozoa rRNA. Bar plots indicating the relative mean abundance per depth for varying taxonomic orders; * indicates undefined taxonomic levels/incertae sedis. Shannon alpha diversity is given as boxplots with whiskers indicating standard deviation across triplicates. Nonmetric multidimensional scaling (NMDS, middle column) ordination plots performed on rRNA contig abundances per sample. Colours indicate different layers and shape different age horizons. Environmental parameters here are given as layer (active layer: AL, transition zone 1: TZ1 and 2: TZ2, and permafrost: PF) as well as age with young soils (AY), 2654–3770-year-old soil of medium age (AM), and old material (AO) of up to 22 100–26 500 years ago.
reads and decreased with depth. Microeukaryotes made up 90.4% of all eukaryotic rRNA contigs and consisted of 58% of the super-group Stramenopiles-Alveolata-Rhizaria (SAR) rRNA contigs (4.7% ± 0.1% of total counts), its relative abundance increasing in the transition zones (Fig. 3C). SAR rRNA contigs, together with Amoebozoa and several Amorphea rRNA contigs made up protozoan microbiome predators.

**Microbiome predatory community**

Out of the 744 rRNA contigs assigned to predatory taxa, 497 were bacterial and 247 eukaryotic. The total proportion of predatory taxa within the total community ranged from 7.9% to 26.3%, peaking at 40–50 cm depth within the permafrost soil first thawed after the collapse. The ratio of protozoa to bacterial predators indicated a higher abundance of protozoa with depth, while predatory bacteria dominated especially within the active layer (Supplementary Table 3). Predator relative abundance significantly correlated with the total (protozoa: F.Model = 5.49, R² = 0.407, P = .006; bacterial predators F.Model = 3.16, R² = 0.283, P = .026) and each subcommunity (Table 2). Several nonpredatory prokaryote abundances decreased in depths where predator abundance increased. Hence, we selected both taxa depleted and enriched in the TZ as potential prey and found that a similar significance level and effect size for Alphaproteobacteria, Actinobacteria, and Bacteroidota as for the total nonpredatory community (Supplementary Table 2). No prey groups could be confirmed statistically, neither by testing of selected taxa (Supplementary Table 2), nor by network-based identification of co-occurrence of predators and prokaryotes, as performed with SPIEC-EASI (Supplementary Fig. 4).

**Bacterial predators**

Bacterial predatory rRNA contigs constituted on average 6.5% ± 0.1% of the total microbial community. They included the former proteobacterial order, now recognised as independent bacterial phyla: Myxococccota, and in lower abundances Bacteriovorax, and as three gammaproteobacterial Lyso bacter and three Vibrionales rRNA contigs (Fig. 3B; Supplementary Table 3). With 48.6% of all predator counts, Myxococccota was the overall most abundant predator with 343 rRNA contigs (6.0% ± 0.1% of total counts).

**Protozoa**

Protozoa included 247 rRNA contigs (5.0% ± 0.1% of total counts), mostly represented by rRNA contigs from the SAR clade. Cercozoa (1.7% ± 0.0%) and Ciliophora (1.7% ± 0.2%) dominated the protozoan community (Fig. 3D; Supplementary Table 3). They were abundant in both transition zones and reached a maximum abundance at 60–70 cm depth (Fig. 3A), the deepest thaw level recorded in 2019 (Fig. 1D). Amoebia were less abundant (0.7% ± 0.0%) and mainly represented in the active layer and TZ1. Lobosa rRNA contigs were most abundant at 40–60 cm within the first thaw horizon TZ1 (Fig. 2D).

**Diversity and drivers of variance within community composition**

Taxonomic richness on average reached 5836 ± 317, while Shannon (S) alpha diversity was on average 7.17 ± 0.53 for the total community (Supplementary Table 1). Both diversity indices indicate overall high alpha diversity in the active layer and a slight decrease with depth (Fig. 3E–H).

The overall community composition most significantly differed with depth (PERMANOVA, F.Model = 6.81, R² = 0.460, P = .001), layer (PERMANOVA, F.Model = 6.42, R² = 0.445, P = .003), and SOM (PERMANOVA, F.Model = 4.77, R² = 0.373, P = .003). Furthermore, H₂O (PERMANOVA, F.Model = 5.23, R² = 0.395, P = .005) explained a significant proportion of the variation of the total community. Age correlated with all community variances except for bacterial predators, while pH was not significant for prokaryotic but for eukaryotic community variance (Table 2). For both, the total and each subcommunity except bacterial predators, only active layer and permafrost differed significantly from each other, while no other layers did (Tukey HSD test, adj. P < .05).

Especially, H₂O, SOM, and layer significantly correlated (ANOVA, P < .005). Variance partitioning indicated that thaw-related edaphic properties (layer, soil moisture), predation (bacterial and eukaryotic predator abundance), and long-term soil characteristics (SOM, pH, age) together explained 51.5% of the variance in beta diversity of the whole community, with the largest partition contributed to thaw (34.8%; Fig. 2). The ordination of subset communities indicated a gradual transition from active layer to TZ1 and TZ2 samples (Fig. 3I–L), while permafrost samples were visually apart as confirmed by PERMANOVA results.

**Discussion**

In this study, we aimed to elucidate the response of the total active permafrost microbial community in up to 26500-year-old material under in situ abrupt thaw conditions. Within the Arctic, particularly Greenland and abrupt erosions sites are highly understudied (Malard and Pearce 2018, Metcalfe et al. 2018). Here, we performed the first total RNA metatranscriptomic analysis of the entire active microbial community in such an environment.

We successfully sequenced soil samples down to depths of ancient, intact permafrost and found increased microbial activity at recently thawed layers and changing abundances of microeukaryotes, including bacteria-feeding protozoa. Using metatranscriptomics, we were able to capture the increased presence and putative activity of fast-growing taxa upon thaw. To understand which taxa are responsible for the increased activity upon thaw, we here analysed the total community for prokaryotic and eukaryotic taxa associated with a potentially copiotrophic as well as predatory lifestyle. In our study, the overall alpha diversity compared to former metatranscriptomic research on permafrost microorganisms (Tveit et al. 2015), or even higher (Hultman et al. 2015, Schostag et al. 2019). The deepest thaw horizon indicated overall fewer and less active taxa than expected, compared to former amplicon-based findings (Scheel et al. 2022), although a few fast-growing taxa (Burkholderiales, Sphingobacterales, and Myxococccota) dominated these depths.

**Community composition in eroding permafrost**

The use of total RNA sequencing enabled the description of the entire soil community with full-length rRNA analysis, including prokaryotic and eukaryotic microbial taxa. The few metatranscriptomic studies from permafrost environments confirmed similar relative abundances of all domains (Hultman et al. 2015, Tveit et al. 2015, Schostag et al. 2019). In our study, most prokaryotes were represented by proteobacterial taxa, which were formerly reviewed as major taxon of both intact permafrost and active layer microbial communities (Jansson et al. 2014, Malard and Pearce 2018). While these studies have detected eukaryotic taxa, we found higher abundances than observed previously (Tveit et al. 2015). The overall low fungal abundances in our study could...
Table 2. Results of permutational multivariate analysis on the mean Bray–Curtis dissimilarities per depths between total community, prokaryotic, bacterial predators (bacpred), microeukaryotic, and protozoan rRNA contigs.

<table>
<thead>
<tr>
<th>Env. parameter</th>
<th>F.Model</th>
<th>R²</th>
<th>F.Model</th>
<th>R²</th>
<th>F.Model</th>
<th>R²</th>
<th>F.Model</th>
<th>R²</th>
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<th>R²</th>
</tr>
</thead>
<tbody>
<tr>
<td>Depth</td>
<td>6.81</td>
<td>0.460</td>
<td>***</td>
<td>6.84</td>
<td>0.461</td>
<td>***</td>
<td>6.44</td>
<td>0.446</td>
<td>***</td>
<td>5.98</td>
</tr>
<tr>
<td>Age</td>
<td>5.11</td>
<td>0.288</td>
<td>*</td>
<td>2.89</td>
<td>0.265</td>
<td>*</td>
<td>2.71</td>
<td>0.253</td>
<td>*</td>
<td>2.91</td>
</tr>
<tr>
<td>Layer</td>
<td>6.42</td>
<td>0.445</td>
<td>***</td>
<td>5.73</td>
<td>0.417</td>
<td>***</td>
<td>5.92</td>
<td>0.425</td>
<td>**</td>
<td>4.12</td>
</tr>
<tr>
<td>AL-PF</td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
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<td></td>
<td></td>
</tr>
<tr>
<td>SOM</td>
<td>4.77</td>
<td>0.373</td>
<td>***</td>
<td>4.76</td>
<td>0.373</td>
<td>***</td>
<td>4.89</td>
<td>0.379</td>
<td>***</td>
<td>3.68</td>
</tr>
<tr>
<td>pH</td>
<td>5.23</td>
<td>0.395</td>
<td>**</td>
<td>5.22</td>
<td>0.395</td>
<td>**</td>
<td>5.42</td>
<td>0.404</td>
<td>**</td>
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</tr>
<tr>
<td>Bacpred&lt;sup&gt;a&lt;/sup&gt;</td>
<td>3.16</td>
<td>0.283</td>
<td>*</td>
<td>3.20</td>
<td>0.286</td>
<td>*</td>
<td>-</td>
<td>-</td>
<td></td>
<td>2.83</td>
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<tr>
<td>Protozoa&lt;sup&gt;b&lt;/sup&gt;</td>
<td>5.49</td>
<td>0.407</td>
<td>**</td>
<td>3.42</td>
<td>0.404</td>
<td>**</td>
<td>5.38</td>
<td>0.402</td>
<td>***</td>
<td>4.64</td>
</tr>
</tbody>
</table>

These were tested against the depth, age, layer, SOM, and moisture (H₂O), and pH. P-values < .05 were considered significant with *P < .05, **P < .01, and ***P < .005. Non-significant (NS) categories and pair-wise contrasts in age or layer were removed. Extended data on predatory and potential prey abundances can be found in Supplementary Table 2. *The effect of bacpred abundance was tested on nonpredatory prokaryotes. **The effect of protozoan abundance was tested on nonpredatory eukaryotes in order to prevent artifacts.

Relate to a rapid decrease of fungi upon thaw (Hultman et al. 2015).

**Thaw and other abiotic drivers**

In this study, abiotic factors, such as the soil depth and layer, SOM content, moisture, and age, as well as biotic processes, such as predation significantly correlated with the overall microbial community. Variance partitioning showed that even a combination of short-, long-term, and predatory effects could only explain half of variance in community composition. Hence, we propose further studies to include the analysis of soil nutrients, including N and phosphorus compounds, and temperature logging before sampling.

The strong correlation between soil moisture and the microbial community can be an artefact as the high soil moisture coincides with the active layer. Here, precipitation is likely the main source of soil moisture, while changes in soil moisture at deeper depths might be most impacted by the melting of the former ice lens. Only the active layer and permafrost communities significantly differed from each other. After the former dispersal-limiting ice lens melted, a slow and gradual migration from the active to deeper thawed layers might be expected (Bottos et al. 2018). Such events could explain why no significant differences in rRNA community composition were found between the two thawed depths. The counts per contig ratios and thus potential activity were lowest in the active layer and maximum in the transition zone samples, where the most abundant contigs were annotated largely annotated as Burkholderiales (Supplementary Fig. 3).

**Copiotrophs in thawing permafrost soil**

Within the prokaryotic community, the most abundant taxa dominated within the recently thawed permafrost material (Supplementary Fig. 3), in agreement with a former amplicon-based study in this site (Scheel et al. 2022). Overall, proteobacterial rRNA contigs were the most abundant and encompassed the most phylogenetically diverse groups. We found that Burkholderiales was the most abundant order, with accordance former DNA-based findings from Arctic soils (Malard and Pearce 2018, Scheel et al. 2022). Gammaproteobacteria include many copiotrophic representatives, which were formerly observed to quickly respond to freshly available resources upon permafrost thaw (Hurst 2019, Schostag et al. 2019). Especially, short-term stressors were seen to trigger a dominance of soil copiotrophs (Koyama et al. 2014, Bang-Andreasen et al. 2020), which might explain why there also had the biggest partition in explaining variance as opposed to long-term and biotic soil properties (Fig. 2).

Burkholderiales peaked in abundance throughout the thawed permafrost depths. Especially Comamonadaceae, a family that consists of many copiotrophic spore-formers, rapidly respond to fresh input of labile nutrients (Fierer et al. 2007, Ho et al. 2017). This suggests that these mineral weathering Burkholderiales (Naylor et al. 2022) might be readily responsive to thawing and hence bioavailable permafrost carbon. The thawing of permafrost soils poses a large potential liberation of organic and inorganic N into the arctic subsols for microbial uptake (Keuper et al. 2012, Voigt et al. 2017, 2020, Maruschak et al. 2021, Wegner et al. 2022). Here, we show a large potential for rapid N-cycling when permafrost thaws. Notably, most contigs with maximum relative abundances at 70 cm, were the abundant Gallionellaceae (Burkholderiales, Gammaproteobacteria) (Hallbeck and Pedersen 2014), which contain nitrate-dependent iron-oxidising taxa (He et al. 2016). This family was previously found in intact ancient permafrost (Alawi et al. 2007, Müller et al. 2018, Scheel et al. 2022).

Similarly, the particularly active Bacteroida order Sphingobacteriales dominated recently thawed permafrost, as found during thaw before (Müller et al. 2018). The relative abundance of Bacteroida increased with depth, potentially indicating that some of its members are more active in permafrost. Yet, high abundances of several copiotrophic order might be an RNA sequencing artefact, where even inactive cells, such as cysts and spores, might contain enough ribosomes so that the contained RNA depicts a corresponding potential activity. This stands against an idea of migrating or being washed down from the active layer, as overall abundances in the active layer were low. In our study site, Sphingobacteriales had the highest relative abundance within the total amplicon-based community composition (Scheel et al. 2022). They then constituted more than half of all counts between 70 and 90 cm depth in 2020. This stands in contrast with very low 16S rRNA-based abundances 1 year earlier (Scheel et al. 2022), which suggests that Sphingobacteriales abundances found in this study reflect a recent increase in response to thaw. Several Bacterioda taxa were documented as initial metabolisers of labile carbon (Fierer et al. 2007, Ho et al. 2017, Stone et al. 2023). Especially at the upper permafrost limit (Müller et al. 2018, Tripathi...
et al. 2018). Bacteroidota abundances also increased in response to thaw (Frank-Fahle et al. 2014, Coolen and Orsi 2015, Deng et al. 2015, Burkert et al. 2019).

**Predation as and biotic driver in response to thaw**

The abundance of both bacterial and protozoan predators correlated with each subcommunity. This supports our findings that indicate that predator abundance was important to the system as a biological control of increased microbial activity but might not have acted as the main driver of taxonomic changes across depths. In a former study investigating the impact of nematodes and protozoa on different bacterial lifestyles under nutrient addition showed that eutrophic taxa often served as prey, the overall bacterial growth being more impacted by nutrient availability than predation (Zelenyev et al. 2006).

**Potential prey and predation interactions**

Several microorganisms have protective measures against predation (Trap et al. 2016). Some protozoan taxa grew less successfully on gram-positive bacteria (Trap et al. 2016), such as Actinobacteriota, which were more abundant in permafrost. This protection might partly be achieved through their formation of filaments, biofilms and colonies, cell size, shape, pigments, and toxins. Here, we noted the decrease of Actinobacteria in the first transition zone, where predators were abundant. Yet, protozoan abundance did not correlate with Actinobacteria abundance more strongly than with other taxa. In contrast, gram-negative bacteria are suitable prey especially to Myxobacteria (the phyla Myxococota and Bdellovibrionota) and protists (Trap et al. 2016). Hence, the increase of activity of gram-negative Proteobacteria in the thaw zone led us to the idea that potentially fast-growing prey could support a consecutive predator succession. Myxobacteria were earlier found to feed select on gram-negative bacteria, while protozoa showed less selective feeding (Zhang and Lueders 2017). This might be reflected in the overall stronger correlation of protozoan abundance with the total and subcommunities. Although species specific feeding had been found among protozoa (Rønn et al. 2001, Pedersen et al. 2011), interkingdom interactions based on a network (Supplementary Fig. 4) indicated more interactions among nonpredatory taxa than among predatory and their potential prey. This could indicate a higher importance of symbiotic processes or competition, as well as other environmental processes that override the potential impact of predation in our system (Fig. 2).

**Bacterial predators dominate shallower thaw layers**

Bacterial predators and particularly Myxococota play key roles in microbial food webs, as recently demonstrated (Dai et al. 2021, Petters et al. 2021). Former studies have shown Myxococota to be abundant in active layer samples (Inglese et al. 2018, Malard and Pearce 2018, Romanowicz and Kling 2022), especially also in this site (Scheel et al. 2022). Within the active layer, seasonally increased prokaryotic biomass could support higher predator abundances. We hypothesize that the spore-containing fruiting bodies in Myxococota (Huntley et al. 2011) might resist the winter freeze-off, while the freeze-induced mortality of prey taxa might supply the growth substrate for Myxococota in spring (Bang-Andreasen et al. 2020). Bdellovibrionota abundances were low. These flagellated bacteria have a smaller prey range and potentially might thrive in competition with Myxococota is lower (Petters et al. 2021).

**Protozoa as active predators present in permafrost soils**

Protozoa of this permafrost study were dominated by Ciliophora, Cercozoa, and Amoebozoa in agreement with former studies (Shatalovich et al. 2009, Schostag et al. 2019). Protozoa were most abundant in both thawed permafrost layers, making them more abundant predators than Myxobacteria, which, in turn, dominated the active and shallow first thaw layer (Fig. 3B and D). The smaller Ciliophora and Cercozoa might have responded quickly to permafrost thaw reaching maximum abundances at both thaw horizons. One potential explanation for the high abundance of protozoa in freshly thawed permafrost could be their mobility in water-saturated soils after the ice lens melted. Further, cyst- and resting-stage-forming protist taxa can respond quickly to environmental changes (Rønn et al. 2012).

Interestingly, among the few permafrost metatranscriptomic studies performed, increases in relative abundance of protozoan taxa with increasing temperature were not mentioned (Schostag et al. 2019), while an increase of Cercozoa was observed (Tveit et al. which could explain their original 2015). Although our relative abundances of Cercozoa matches those of former studies (Geisen et al. 2015, Tveit et al. 2015, Schostag et al. 2019), we expected higher overall abundances at the most recently available thaw horizon at 90 cm. Due to their relatively smaller size, Cercozoa reproduce faster than larger microeukaryotes (Rønn et al. 2012). Furthermore, members of this group have been shown to respond to permafrost thaw within days (Schostag et al. 2019), as well as to temperature increases in Arctic peat (Tveit et al. 2015). Ciliates and Amoeba respond very rapidly to changes in soil moisture (Coleman and Wall 2015) and temperature (Schostag et al. 2019) and have even been found in ancient Siberian permafrost (Shatalovich et al. 2009). Although some studies reported living nematodes and Amoebozoa in ancient permafrost (Shmakova and Rivkina 2015, Shatalovich et al. 2023), their response to permafrost thaw remains unclear.

Our findings suggest that eukaryotic predators dominate during the initial stages of the thaw-triggered microbial activity, while during later stages of thaw increased relative abundance of the less mobile bacterial predators occurred. Furthermore, protozoans might also benefit from the unselective feeding mechanisms in contrast to the more selective Myxobacteria (Zhang and Lueders 2017). This potentially explains why we observed changes in distribution with depth and significant impact of protozoan, but not myxobacterial abundance on potential prey taxa.

**Microbial activity during thaw and potential coalescence**

Predation bears the obvious potential to reduce and hence control microbial population growth, but also has been found to lead to higher yields in artificial diverse incubation communities. This effect might be due to reduced competition among prey taxa (Saleem et al. 2012), through that keeping populations in growth phase (Mattison and Harayama 2001), as well as potentially leaching nutrients from prey cell lysis. This could explain, why predator abundance correlated with every subcommunity, likely indicating indirect rather than direct predation effects on the whole community. We noted that both prokaryotic and eukaryotic taxa were present among the most active contigs in the thawed layers. Thus, our findings of higher RNA:DNA ratios at the freshly thawed 50 and 80 cm depths indicated higher ribosomal relative abundance of both domains at these depths (Table 1, Supplementary Fig. 1).

The difference in the composition of the active microbial community between shallow and deeper thawed permafrost lay-
ers could result from multiple ecological processes simultaneously impacting the microbial community composition. For example, (1) a steady migration of microorganisms from the active layer to the thawed layers, compared to the below permafrost layer, and (2) rapidly increased microbial activity in recently thawed, former permafrost layers. Ecological responses of soil microbiomes to environmental stress are nonlinear and complex, arising from simultaneous stochastic and deterministic environmental selection (Doherty et al. 2020, Ernakovich et al. 2022).

The peaking activity in ancient, former frozen permafrost, could stem from endemic spores and resting stages. Situated by the Zackenberg river, the oldest sand and rubble deposits (Ao) indicate a fluvial origin, potentially as part of the Zackenberg river or glacial meltwater delta into the fjord roughly 30 000 years ago (Gibert et al. 2017), while younger material could resemble a former pond, that later became overgrown (AM). Arctic fluvial habitats showed higher abundances of Bacteroida and Proteobacteria in a summer glacial river, but also the Arctic Lake Hazen sediments (Cavaco et al. 2019) and subglacial habitats (Achberger et al. 2017), which could explain their origin even in our samples of ancient and still intact permafrost. Simultaneously, dominance of Bacteroida was also found in permafrost metatranscriptomes during thaw (Hultman et al. 2015, Schostag et al. 2019).

In contrast to the idea of increased activity in former cold-adapted taxa, earlier findings showed the potential migration of taxa from the active layer to thawing permafrost (Rillig et al. 2015). In their previous incubation study, Monteux and colleagues found an increase of Bacteroida abundance upon thaw, after inoculating permafrost with grassland soil (Monteux et al. 2020). Both the processes of a thaw control as well as implantation of temperate soil have resulted in elevated nitrification and CO₂ emissions (Monteux et al. 2020). This highlights the importance of understanding community dynamics in response to thaw. This process was also observed upon transplantation of grassland soil microbiomes onto permafrost (Monteux et al. 2020). Such mixing conditions occurs naturally during cryoturbation of permafrost, such as during a collapse (Gittel et al. 2014, Schnecker et al. 2014), active layer detachments (Inglese et al. 2018), or increased root growth (Monteux et al. 2020). Although we could not statistically separate the potential effects of isolated permafrost thaw and coalescence with the active layer community, they might supply a higher functional diversity to overcome this limitation and hence enhance C and N from thawed permafrost.

**Conclusion**

We have described the total prokaryotic and microeukaryotic community composition abrupt permafrost erosion stress in High Arctic Greenland based on total RNA sequencing of in situ samples. The composition of the total community correlated both with short- and long-term soil properties, including thaw layer, soil moisture, and organic matter content. The high microbial activity we found within freshly thawed permafrost, as indicated by large increases in the RNA:DNA ratio, coincided with an increase of copiotrophic taxa, including representation of Spinogobacteriales and Burkholderiales. This potentially rapid development of microbial biomass at thawed depths also formed the base of a solid bacteria-feeding community dominated by protozoa in recently thawed soil and bacterial predators just below the historical active layer boundary. We found that predation, addi-

**Author contributions**

Maria Scheel (Conceptualization, Data curation, Formal analysis, Investigation, Methodology, Project administration, Resources, Software, Visualization, Writing—original draft, Writing—review & editing), Athanasios Zervas (Data curation, Investigation, Methodology, Software, Supervision, Writing—review & editing), Ruud Rijkers (Formal analysis, Methodology, Software, Validation, Writing—review & editing), Alexander Tysdal Tveit (Conceptualization, Investigation, Methodology, Validation, Writing—review & editing), Flemming Ekeland (Conceptualization, Validation, Writing—review & editing), Francisco Campuzano Jiménez (Formal analysis, Software, Visualization, Writing—review & editing), Torben Røjle Christensen (Data curation, Funding acquisition, Project administration, Resources, Supervision, Writing—review & editing), and Carsten Suhr Jacobsen (Conceptualization, Funding acquisition, Methodology, Project administration, Resources, Supervision, Writing—review & editing).

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**Supplementary data**

Supplementary data is available at FEMSEC Journal online.

**Conflict of interest** None declared.

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**Data availability**

The raw sequence data of this study were deposited in the NCBI Sequence Read Archive and can be accessed through accession number PRJNA939404.

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