



## Exposure to Air Pollution inside Electric and Diesel-Powered Passenger Trains

Andersen, Maria Helena G.; Johannesson, Sandra; Fonseca, Ana Sofia; Clausen, Per Axel; Saber, Anne Thoustrup; Roursgaard, Martin; Loeschner, Katrin; Koponen, Ismo K.; Loft, Steffen; Vogel, Ulla; Møller, Peter

*Published in:*  
Environmental Science & Technology

*DOI:*  
[10.1021/acs.est.8b06980](https://doi.org/10.1021/acs.est.8b06980)

*Publication date:*  
2019

*Document version*  
Peer reviewed version

*Citation for published version (APA):*  
Andersen, M. H. G., Johannesson, S., Fonseca, A. S., Clausen, P. A., Saber, A. T., Roursgaard, M., ... Møller, P. (2019). Exposure to Air Pollution inside Electric and Diesel-Powered Passenger Trains. *Environmental Science & Technology*, 53(8), 4579-4587. <https://doi.org/10.1021/acs.est.8b06980>

1 **Exposure to air pollution inside electric and diesel-powered passenger trains**

2 Maria Helena G. Andersen<sup>1,2\*</sup>, Sandra Johannesson<sup>3</sup>, Ana Sofia Fonseca<sup>2</sup>, Per Axel Clausen<sup>2</sup>,  
3 Anne Thoustrup Saber<sup>2</sup>, Martin Roursgaard<sup>1</sup>, Katrin Loeschner<sup>4</sup>, Ismo K. Koponen<sup>2</sup>, Steffen Loft<sup>1</sup>,  
4 Ulla Vogel<sup>2,5</sup>, Peter Møller<sup>1</sup>

5

6 <sup>1</sup>Department of Public Health, Section of Environmental Health, University of Copenhagen, Øster  
7 Farimagsgade 5A, DK-1014 Copenhagen K, Denmark

8 <sup>2</sup>The National Research Centre for the Working Environment, Lersø Parkalle 105, DK-2100  
9 Copenhagen Ø, Denmark.

10 <sup>3</sup>Department of Occupational and Environmental Medicine, Sahlgrenska Academy at University  
11 of Gothenburg, 40530 Gothenburg, Sweden.

12 <sup>4</sup>National Food Institute, Technical University of Denmark, DK-2800 Kgs. Lyngby, Denmark

13 <sup>5</sup>Department of Health Technology, Technical University of Denmark, DK-2800 Kgs. Lyngby,  
14 Denmark

15 \*Corresponding author: mhar@sund.ku.dk; mga@nfa.dk

16

17 **Keywords**

18 Diesel exhaust, diesel train, personal exposure, ultrafine particles, black carbon, nitrogen oxides,  
19 underground station, commuter

20

21

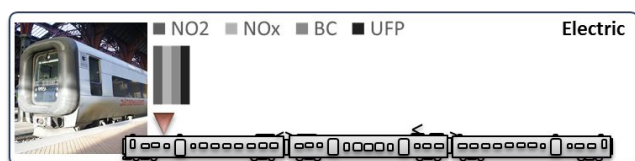
22

23

24

25

## 26 Table of contents/Abstract art



27

28

## 29 Abstract

30 Diesel-powered trains are used worldwide for passenger transport. The present study aimed to  
31 assess air pollution concentrations in passenger cars from diesel and electric trains. Personal  
32 exposure monitoring (6-7 hours per day) was carried out for 49 days on diesel and 22 days on  
33 electric trains. Diesel trains had higher concentrations of all the assessed air pollution components.  
34 Average increases (and fold differences) in passenger cars of diesel trains compared with electric  
35 trains were for ultrafine particles 212,000 particles/cm<sup>3</sup> (35-fold), black carbon 8.3 µg/m<sup>3</sup> (6-fold),  
36 NO<sub>x</sub> 316 µg/m<sup>3</sup> (8-fold), NO<sub>2</sub> 38 µg/m<sup>3</sup> (3-fold), PM<sub>2.5</sub> 34 µg/m<sup>3</sup> (2-fold) and benzo(a)pyrene 0.14  
37 ng/m<sup>3</sup> (6-fold). From time-series data, the pull and push movement modes, the engine in use and  
38 the distance to the locomotive influenced the concentrations inside the diesel trains. In conclusion,  
39 concentrations of all air pollutants were significantly elevated in passenger cars in diesel trains  
40 compared to electric trains.

41

## 42 Introduction

43 Railway passenger and freight transport activity have been growing globally in the last 40 years<sup>1</sup>.  
44 Electricity is the power supply of choice in trains in many areas, but a significant proportion of  
45 commuter and freight trains worldwide are still powered by diesel fuel. It has been shown that

46 diesel-powered trains increase the concentrations of fine particulate matter (PM<sub>2.5</sub>) and ultrafine  
47 particles (UFP; <0.1 μm) in the open air close to the railway<sup>2, 3</sup>. However, PM has been shown to  
48 also penetrate into passenger cars of the train<sup>4, 5</sup>. A review of commuter exposure to particles  
49 reported three times higher UFP concentrations in passenger cars of diesel than electric trains,  
50 with highest concentrations when diesel-propelled locomotives operate in pull mode<sup>6</sup>. Two more  
51 recent studies also reported significant differences of UFP concentrations inside passenger cars of  
52 diesel trains depending on position of locomotive<sup>4, 7</sup>. In general, high variability of UFP  
53 concentrations has been reported inside diesel trains. More knowledge on seasonal variation and  
54 other components of the complex diesel exhaust mixture penetrating to train cars is needed.  
55 Electric trains generate only particle emissions from non-exhaust sources such as wear and  
56 mechanical friction on rail tracks, wheels and brakes, as well as friction and arcing of the electrical  
57 wires. The majority of the studies on electric trains have addressed particle exposure on train  
58 platforms and tunnels<sup>8</sup>, although a few recent studies assessed exposures and factors affecting  
59 exposures inside passenger cars<sup>5, 9</sup>.

60

61 High concentrations of diesel exhaust in passenger cars are worrying because it is associated with  
62 increased risk of lung cancer as well as other lung and cardiovascular diseases<sup>10-12</sup>. In addition to  
63 passengers, train conductors, engine drivers, maintenance staff, platform staff and other train  
64 personnel can be exposed. Our study was initiated because of concern about potential health  
65 effects in workers and passengers who commute in diesel-powered trains. In Copenhagen,  
66 Denmark, the urban train system is electrified, but co-exists with diesel-powered trains that travel  
67 further on intercity routes. A recent report showed that two types of diesel rolling stock used in  
68 Denmark gave rise to higher UFP exposure to staff working in the passenger cars than what would  
69 occur on the most busy street in Copenhagen<sup>13</sup>. The aim of the present study was to perform a

70 comprehensive comparison of the concentration levels inside passenger cars of diesel-powered  
71 and electric trains including both particulate and gaseous air pollutants.

72

## 73 **Methods**

### 74 *Study design*

75 Twenty-nine volunteers were recruited to sit inside the front car of Danish State Railways trains  
76 running in the Zealand region, carrying portable measuring devices and samplers. The  
77 concentrations characterization was performed in the same type of trains on three consecutive days  
78 each week (Tuesday, Wednesday and Thursday) for 6-7 hours per day. A pilot study was  
79 conducted during two days in March in order to characterize the size distributions, test the  
80 concentration contrast and test the chosen routes. The main study period covered a total of 23  
81 weeks from May to December 2017. Out of the 23 monitoring weeks, seven were conducted in  
82 electric trains and 16 in diesel trains. In addition to measurements in trains, two more days of  
83 measurements were done on a busy multimodal underground train platform in Copenhagen, as a  
84 worst-case scenario used opportunistically to assess the contribution of the studied diesel trains in  
85 the underground environment.

86

### 87 *Trains and train routes*

88 The electric trains operated on a route from Østerport station to Elsinore station (46 km). The train  
89 route was taken forward and back 3 times per day, for a total of 5-hours and 37 minutes per day,  
90 including waiting time at Østerport and Elsinore stations, starting at 9h39 and ending at 15h16.  
91 The rolling stock is termed Litra ET and runs on alternating current in overhead power lines (25  
92 kV and 50 Hz).

93

94 As the diesel train exposure scenario, a double deck train propelled by a diesel locomotive (Litra  
95 ME, from 1981-85, 33 units in circulation) was chosen in a route from Østerport to Kalundborg  
96 station (114 km), return to Copenhagen Central Station, and from there to Holbæk (part of  
97 Kalundborg line, 68 km) and back to Copenhagen Central Station. The route took a total of 7  
98 hours per day, including waiting time at the stations, starting at 8h34 and ending at 15h32. This  
99 route was chosen because the same locomotive was used all the day (defined as “diesel A”). For  
100 the measurements in October and November, the trip was firstly heading to Holbæk and then to  
101 Kalundborg, driven with three different ME locomotives and saving one hour of waiting time at  
102 the station (defined as “diesel B”). The driving locomotive code was noted every day for the diesel  
103 scenario. The diesel routes include a passage at one underground station in the Copenhagen city  
104 centre (Nørreport Station). The plan for the different weeks and schedules are available in  
105 supporting information (SI, Tables S1 and S2).

106

107 The windows in both passenger train cars (electric and diesel) were sealed and could not be  
108 opened. Furthermore, the ventilation system was not passenger-adjustable.

109

### 110 ***Underground station***

111 The underground train platform at Nørreport station is a 200 m long single platform positioned  
112 between two tracks. It is equipped with a modern ventilation system (from 2014) and serves both  
113 electric and diesel regional and intercity trains. Measurements were carried out on two different  
114 days: on the first day (Friday January 26<sup>th</sup>) all train types were running as usual, but on the second  
115 day (Friday February 2<sup>nd</sup>) all ME locomotives had been taken out of circulation for maintenance,  
116 providing a unique opportunity to perform measurements at the underground station both with and  
117 without the ME-diesel locomotives. The monitoring instruments were maintained in the same

118 position on the train platform for 4 hours. The train passages were noted for both tracks in both  
119 days.

120

### 121 *Measurements, instrumentation and chemical analyses*

122 All air pollution measurements were performed as personal monitoring with the volunteers  
123 carrying the equipment. In the diesel trains the volunteers were sitting in the compartment in the  
124 front section of the first passenger car (Figure S2) and in the electric trains they were sitting in the  
125 first passenger car. Black carbon mass concentration (BC) and UFP number concentration were  
126 measured on each monitoring day during the total period of the study, both on-board of the diesel  
127 and electric trains and on the platform of the underground station. Nitrogen oxides (NO<sub>2</sub> and NO<sub>x</sub>;  
128 whole period), aldehydes (10 weeks from May to September), PM<sub>2.5</sub> and polycyclic aromatic  
129 hydrocarbons (PAHs; 9 weeks in October and November) were sampled on-board of the electric  
130 and diesel trains (Table S1). Additionally, PM was collected by means of a mid-volume  
131 electrostatic sampler, described elsewhere<sup>14</sup>, in the compartment in the front train car during the  
132 monitoring weeks in December. Measurements to investigate concentration gradient inside the  
133 trains were performed for three days, where one person sat in the compartment in the front section  
134 of the first passenger car, while another person sat each day at a different position in the lower  
135 deck of the train (for entire day trip), each of them carrying one BC and one UFP measuring  
136 device. A summary of the measurements, instruments, periods and sites monitored as well as time  
137 resolutions and flows used is presented in SI (Table S3).

138

### 139 *Black carbon*

140 Black carbon was measured with 1-minute resolution using two MicroAeth AE51 aethalometers  
141 (Aethlabs, San Francisco, CA) with high degree of comparability (correlation coefficient,  $r=0.95$ ).

142 The filters inside the aethalometer were changed after each monitoring day in the diesel trains, and  
143 after the three monitoring days in the electric trains.

144

#### 145 *Ultrafine particles*

146 UFP concentrations were measured using both diffusion chargers DiSCmini (DM; DiSCmini,  
147 Matter Aerosol AG, Wohlen, Switzerland) and NanoTracer (NT; Aerasense NanoTracer, Oxility,  
148 Eindhoven, the Netherlands). For the two-day pilot study the NanoScan SMPS model 3910 (TSI,  
149 Shoreview, MN, USA) was also used. The devices were synchronized with a working computer's  
150 clock. Two DM devices were used for the gradient measurements with high degree of  
151 comparability ( $r=0,96$ ). Measurement time resolutions were 1-second for DM, 16-seconds for NT  
152 and 1-minute for NanoScan SMPS (Table S3).

153

#### 154 *Nitrogen oxides*

155 Passive Ogawa samplers (Ogawa, FL, USA) were used for personal sampling of nitrogen oxides.  
156 Two samplers were used over the three consecutive days in each exposure scenario and kept in  
157 sealed plastic bags overnight at 4°C (and after completed sampling). Samplers were analysed by  
158 ion chromatography and corrected with field blanks<sup>15</sup>. The limit of detection (LOD) for the  
159 sampling time of 18-21 hours was 0.7 µg/m<sup>3</sup> for NO<sub>2</sub> and 1.4 µg/m<sup>3</sup> for NO<sub>x</sub>. None of the samples  
160 were below the LOD.

161

#### 162 *Aldehydes*

163 Aldehydes were measured with Sep-Pak XPoSure aldehyde samplers (Waters, MA, USA)  
164 connected to pumps. The samplers were closed with caps and kept overnight at 4°C during the  
165 three cumulated sampling days for each exposure scenario and in the end kept at -20°C. The  
166 samplers were analysed for 13 different aldehydes and one ketone using high-performance liquid



167 chromatography with UV detection. The sampled air volume was 0.21-0.27 m<sup>3</sup> and the LOD was  
168 0.5 µg/sample for each substance.

169

#### 170 *PM<sub>2.5</sub> and polycyclic aromatic hydrocarbons*

171 PM<sub>2.5</sub> was collected on filters (Teflon w/ring PALL 2.0 µm, 37 mm) using a cyclone (Triplex  
172 cyclone, BGI, USA) at 1.5 L/min. A sequence set of two XAD-2 tubes (sample + backup tube) for  
173 sampling of gaseous PAHs were attached after the cyclone (Figure S1). After the sampling week  
174 was completed, the XAD-2 tubes were wrapped in aluminium foil and kept at -20°C. The filter  
175 cassettes were wrapped in aluminium foil and kept in the dark, dry and at room temperature before  
176 and after use. The Teflon filters were weighed before and after sampling in controlled climate  
177 conditions. LOD for particle mass was 29 µg. Two filters were excluded due to sampling failure  
178 and/or failure in the chemical analysis.

179

180 Particles on filters were extracted and analysed for the content of 16 US EPA priority PAHs  
181 (Table S5) using high-resolution gas chromatography/low-resolution mass spectrometry<sup>16</sup>. Field  
182 blanks were analysed along with the samples, and certified reference material was used for quality  
183 control. The XAD-2 from the sample and backup tube were combined and sonicated with 5 mL  
184 cyclohexane for 60 minutes and analyzed by gas chromatography-mass spectrometry for the same  
185 PAHs as in filters, analyzed in a different laboratory as previously described<sup>17</sup>. LODs for gaseous  
186 PAHs were 0.1-0.6 pg/µL extract depending on the individual PAH.

187

#### 188 *PM metal content*

189 PM collected with the electrostatic sampler in weeks 23 and 24 in diesel A scenario was analysed  
190 for the content of 24 elements by inductively coupled plasma mass spectrometry. Detailed  
191 information is available in SI, Table S6.

192

### 193 *Time-series data analyses*

194 The time-series data from BC, DM and NT (also from NanoScan, for the pilot study) were  
195 synchronized in start and ending times and grouped by scenario (diesel A, diesel B and electric).  
196 The BC data were collected with two duplicate devices, which were averaged for each measured  
197 resolution unit (minute). Data for each monitoring day from each device were aggregated in 10  
198 minutes intervals and the time slots of the different days were averaged. The averages and  
199 percentiles for each 10-minute average concentration over time of the day within each scenario  
200 were then plotted, as well as 10-minute maximum values using R statistical language<sup>18</sup>. The BC  
201 and DM time-series from the concentration gradient measurements and from the underground  
202 station were kept and re-sampled in the smallest time resolution possible for comparison (i.e. 1  
203 minute). Some of the monitoring days were excluded from the analysis (Table S4).

204

### 205 *Statistics*

206 The Welch t-test was used to compare the concentrations in diesel and electric scenarios using R  
207 statistical language<sup>18</sup>. The average of daily 10-minute slots for all days measured was used for  
208 calculating the mean, SD and range for time-series data from BC and UFP. The average  
209 concentration of the two duplicate personal samplers used per week was used for calculating the  
210 mean concentrations, SD and the range in concentrations for NO<sub>x</sub>, NO<sub>2</sub>, aldehydes, PM<sub>2.5</sub> and the  
211 PAHs.

212

## 213 **Results**

### 214 *On-board the trains*

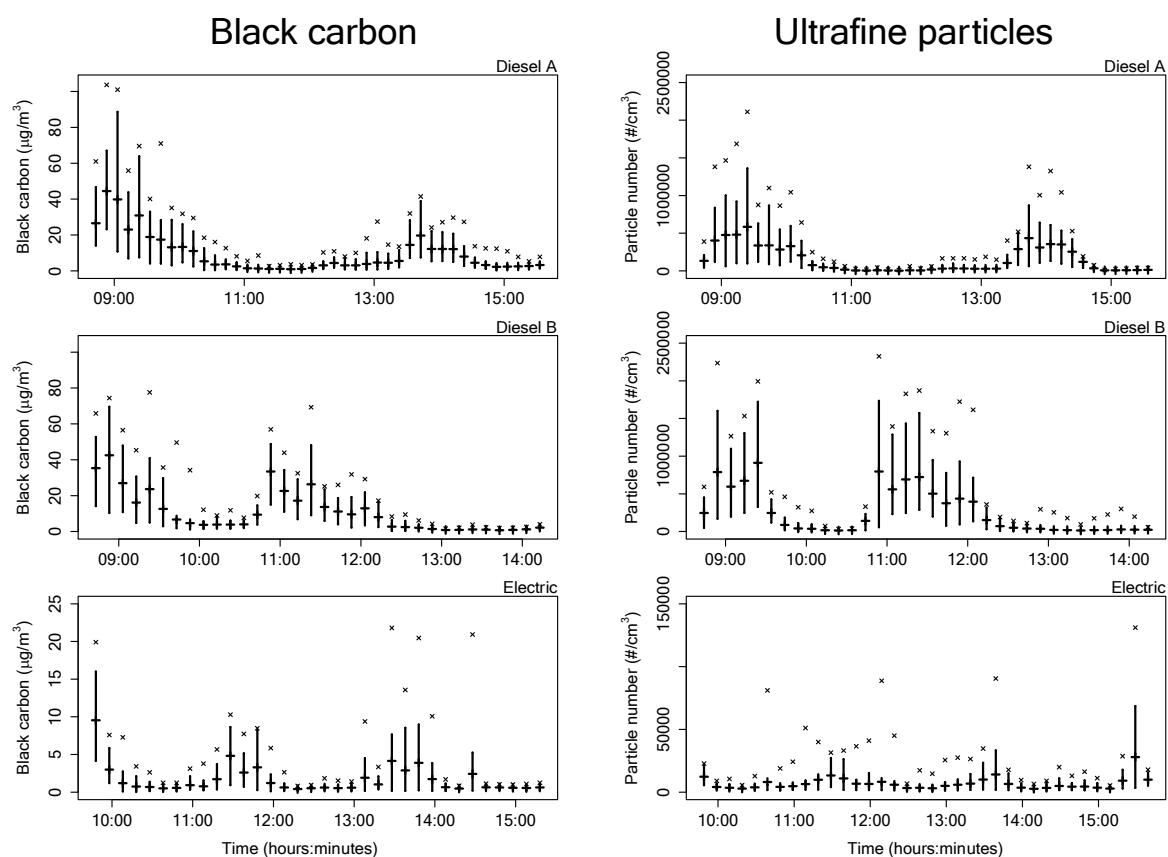
#### 215 *Black carbon and ultrafine particles*

216 Total particle number concentration and average size of UFP were measured in a pilot study in  
217 Diesel A scenario with three different devices, showing similar readings of particle number  
218 concentrations between devices (Figure S3). Detailed output figures from the NanoScan SMPS  
219 device from both electric and diesel pilot measurements are in SI (Figures S4 and S5). For  
220 personal monitoring in the main study the NanoScan device was not used since it is less portable.

221

222 Figure 1 shows the mean daily particle concentration patterns inside the train in the Diesel A,  
223 Diesel B and Electric exposure scenarios. The patterns for BC mass concentrations and UFP  
224 number concentrations (measured by DM equipment) showed similar trends (with daily averages  
225 from both devices with  $r=0.75$ ). The NT measurements were similar to the DM data and are  
226 shown in SI (Figure S6). For both BC and UFP in the Diesel A scenario, a large increase in  
227 concentrations was observed between 8h30 and 10h, and a small increase between 13h30 and  
228 14h30. A similar trend was observed for the Diesel B scenario, with the second peak appearing  
229 earlier in the day. The increased concentrations of BC and UFP in both diesel scenarios occurred  
230 when the locomotive was pulling the train (leaving Copenhagen) and the lower concentrations  
231 when the locomotive was pushing the train (returning to Copenhagen). When the locomotive was  
232 in front, higher concentrations of BC and UFP were measured inside the passenger car with  
233 average particle number concentrations of around 400,000 particles/cm<sup>3</sup> and BC mass  
234 concentration exceeding 20 µg/m<sup>3</sup>. For the 10-minutes average maximums, the UFP  
235 concentrations exceeded 2 million particles/cm<sup>3</sup> and BC mass concentration was 70-100 µg/m<sup>3</sup>  
236 (Figure 1). When the locomotive was in push mode the averaged concentrations were 10,000-  
237 30,000 particles/cm<sup>3</sup> and 2-3 µg/m<sup>3</sup> BC. The concentrations of BC and UFP inside the electric  
238 trains were considerably lower than in the diesel trains. A small increase in concentrations could  
239 be observed when the train was at Østerport station in Copenhagen, in the beginning and end of  
240 the measurements, and also in the time intervals 11h16-11h39 and 13h16-13h39 when the

241 volunteers waited on the Østerport station platform to change trains. At the terminal station for  
 242 electric trains in Elsinore, the concentrations were at the same level as when the train was in  
 243 movement (between 3,000-7,000 particles/cm<sup>3</sup> and 0.5-4 µg/m<sup>3</sup> BC), while at Østerport station,  
 244 shared with diesel trains, the averaged concentrations were 10,000-13,000 particles/cm<sup>3</sup> and 3-5  
 245 µg/m<sup>3</sup> BC. Table 1 summarizes the daily averaged concentrations of BC and UFP for diesel (both  
 246 diesel A and B) versus electric scenario and Table 2 shows the differences between movement  
 247 mode (pull/push, forward/backward, for diesel and electric) and for the different train scenarios  
 248 (diesel A, diesel B and electric). For BC, concentrations were 6-fold higher in the diesel compared  
 249 to the electric scenario and for UFP 35- and 20-fold higher (measured with DM and NT,  
 250 respectively).  
 251



252  
 253 Figure 1 - Daily temporal BC mass concentrations and UFP number concentrations (measured by  
 254 DiscMini) in the defined scenarios (Diesel A, Diesel B and Electric). The symbols represent the

255 arithmetic mean (horizontal lines), the 10<sup>th</sup> to 90<sup>th</sup> percentile distributions (vertical lines), and the  
 256 maximum values (point crosses) for the 10 minutes averaged data. BC data were collected from 28  
 257 monitoring days for the Diesel A scenario, for 17 days for the Diesel B scenario and 18 days for  
 258 the Electric scenario. UFP data were collected from 21 days for Diesel A, 18 days for Diesel B  
 259 and 17 days for Electric (Table S4 presents missing data). Note the different y-axis ranges for  
 260 diesel and electric trains.

261

262 Table 1 – Black carbon, ultrafine particles, nitrogen oxides, aldehydes, PM<sub>2.5</sub> mass concentration  
 263 and PAHs in particles mass and gas phase, respectively.

	Diesel train			Electric train			Fold difference (diesel vs electric)	Mean difference (CI 95%)
	n	Mean±SD <sup>1</sup>	Range <sup>1</sup>	n	Mean±SD <sup>1</sup>	Range <sup>1</sup>		
Black carbon (µg/m <sup>3</sup> )	21	10.0 ± 3.3	3.8-15.1	7	1.7 ± 0.6	1.2-3.2	6	8.3 (7.3; 9.3)***
UFP from DM (#/cm <sup>3</sup> )	15	218,000 ± 140,000	31,700-560,000	6	6,150 ± 2,310	4,100-12,700	35	212,000 (167,000; 257,000)***
UFP from NT (#/cm <sup>3</sup> )	17	153,000 ± 81,000	29,200-425,000	7	7,760 ± 4,200	3,670-22,600	20	145,000 (121,000; 169,000)***
NO <sub>x</sub> (µg/m <sup>3</sup> )	14	364 ± 81	275-531	7	48 ± 16	25-65	8	316 (268; 365)***
NO <sub>2</sub> (µg/m <sup>3</sup> )	14	53 ± 16	35-99	7	16 ± 10	6-31	3	38 (26; 49)***
Aldehydes (sum) (µg/m <sup>3</sup> )	6	50 ± 8	37-59	4	41 ± 2	39-44	1.2	NC <sup>2</sup>
Formaldehyde (µg/m <sup>3</sup> )	6	9.0 ± 2.1	6.4-11.4	4	5.2 ± 0.3	4.8-5.5	1.7	NC <sup>2</sup>
PM <sub>2.5</sub> (µg/m <sup>3</sup> )	6	68 ± 15	51-96	3	32 ± 7	24-37	2	NC <sup>2</sup>
ΣPAH in PM <sub>2.5</sub> (ng/m <sup>3</sup> )	6	1.7 ± 0.6	1.0-2.5	3	0.31 ± 0.13	0.21-0.45	6	NC <sup>2</sup>
BaP (ng/m <sup>3</sup> ) in PM <sub>2.5</sub>	6	0.17 ± 0.09	0.08-0.30	3	0.03 ± 0.003	0.028-0.034	6	NC <sup>2</sup>
Σ PAH XAD-2 tubes (ng/m <sup>3</sup> )	6	165 ± 56	110-261	3	146 ± 48	116-201	1.1	NC <sup>2</sup>
Naphthalene XAD-2 tubes	6	134 ± 26	99-170	3	108 ± 55	67-170	1.2	NC <sup>2</sup>

264 (ng/m<sup>3</sup>)  
 265 n, number of weeks sampled (3 days per week); SD, standard deviation; CI, confidence interval;  
 266 UFP, ultrafine particles; DM, DiscMini; NC, not calculated; NT, NanoTracer;  $\Sigma$ PAH, total  
 267 amount of PAH in particles collected on filter or collected in the XAD-2 tubes; BaP,  
 268 benzo(a)pyrene. <sup>1</sup>Mean and Range determined from values over daily averages of 10-minute time  
 269 periods; <sup>2</sup>Limited number of samples to calculate a reliable mean difference. Diesel trains include  
 270 both scenarios A and B; \*\*\* p<0.001

271 Table 2 – Black carbon and ultrafine particle concentrations for movement mode and scenario

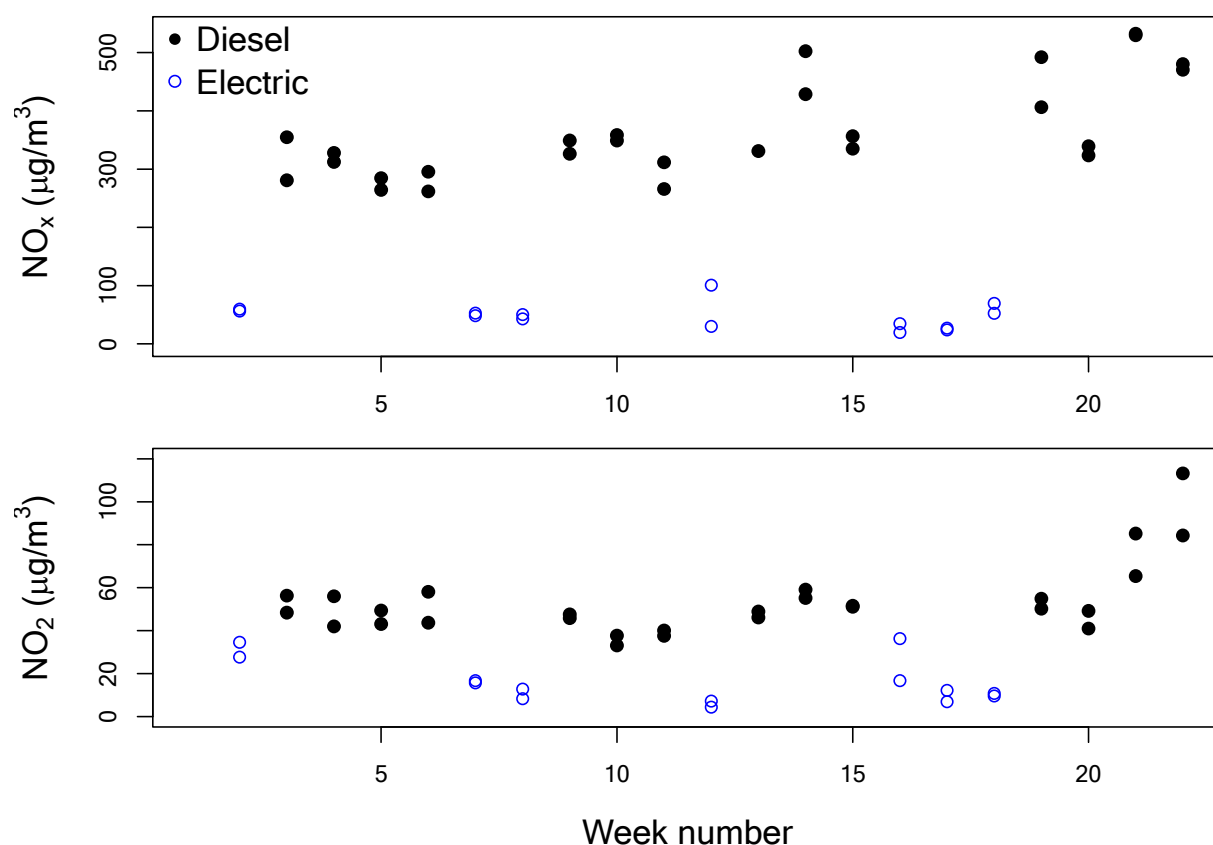
		Pull mode/forward		Push mode/backward		Fold difference (pull vs push)	Mean difference (CI 95%)
		Mean±SD <sup>1</sup>	Range <sup>1</sup>	Mean±SD <sup>1</sup>	Range <sup>1</sup>		
Scenario A	Black carbon (µg/m <sup>3</sup> )	19.8 ± 10.7	8.0-44.5	2.0 ± 0.9	0.9-3.3	10	17.8 (12.1; 23.5)***
	UFP from DM (#/cm <sup>3</sup> )	347,600 ± 111,200	132,400-583,500	10,400 ± 7,800	3,300-34,000	33	337,200 (278,000; 396,600)***
	UFP from NT (#/cm <sup>3</sup> )	279,500 ± 88,000	128,000-500,200	13,100 ± 12,000	5,200-53,800	21	266,400 (219,000; 313,700)***
Scenario B	Black carbon (µg/m <sup>3</sup> )	20.8 ± 10.4	8.0-42.5	2.5 ± 1.8	0.7-6.7	8	18.3 (12.5; 24.2)***
	UFP from DM (#/cm <sup>3</sup> )	539,600 ± 228,000	152,400-912,300	27,600 ± 18,700	13,100-87,400	20	512,000 (385,500; 638,600)***
	UFP from NT (#/cm <sup>3</sup> )	400,400 ± 130,800	185,200-656,300	26,000 ± 23,000	11,500-103,600	15	374,400 (301,200; 447,500)***
Electric	Black carbon (µg/m <sup>3</sup> )	1.1 ± 0.8	0.4-3.0	0.9 ± 0.5	0.5-1.9	1.2	0.3 (0.4; 0.9)
	UFP from DM (#/cm <sup>3</sup> )	4,500 ± 1,900	2,700-8,300	5,500 ± 2,200	2,800-9,900	0.8	960 (900; 2,900)
	UFP from NT (#/cm <sup>3</sup> )	5,200 ± 1,400	4,000-8,000	7,400 ± 5,700	3,900-24,000	0.7	2,200 (1,500; 5,900)

272 SD, standard deviation; CI, confidence interval; UFP, ultrafine particles; DM, DiscMini; NT,  
273 NanoTracer; <sup>1</sup>Mean and Range determined from values over daily averages of 10-minutes data  
274 collected with trains in movement, excluding time at station. The 10-minute resolution averages  
275 were not completely synchronized with the real position of the train, but buffered with time at  
276 station; \*\*\* p<0.001

277

### 278 Nitrogen oxides

279 NO<sub>x</sub> on board the diesel trains was nearly 8 times higher than on the electric trains (Table 1 and  
280 Figure 2). NO<sub>2</sub> concentrations were about 3-fold higher on the diesel trains compared with the  
281 electric trains. There was a trend of increasing concentrations of NO<sub>2</sub> during the two last weeks of  
282 the project (end of November).



283

284 Figure 2 - NO<sub>x</sub> (upper) and NO<sub>2</sub> (lower) air concentrations on board the diesel and the electric  
285 trains. Each point represents the average air concentration over the three consecutive days,

286 collected with two personal samplers per week. Week numbers are reference numbers for dates  
287 presented in supporting information (Table S1)

288

#### 289 *Aldehydes*

290 The concentrations of aldehydes were very low, and therefore these samplings were discontinued  
291 after 10 weeks. For eight of the 13 analysed aldehydes (and acetone), all or the majority of the  
292 samples were below the LOD. The other six (acetaldehyde, acetone, decanal, formaldehyde,  
293 hexanal and nonanal) are presented as aldehyde total sum (Table 1). The mean concentration of  
294 formaldehyde was 1.7-fold higher in the diesel than electric trains (Table 1).

295

#### 296 *PM<sub>2.5</sub> and polycyclic aromatic hydrocarbons*

297 Five filter samples had mass levels below LOD (all corresponded to the electric scenario). The  
298 mean concentrations of PM<sub>2.5</sub> inside the diesel trains were 2-fold higher than inside the electric  
299 trains. Concentrations of benzo(a)pyrene (BaP) and the sum of particle-bound PAHs were 6-fold  
300 higher in diesel than in electric trains (Table 1). Among the 16 PAHs analysed in the particle  
301 mass, seven were quantified in all filters from both diesel and electric trains (pyrene,  
302 benz(a)anthracene, benzo(b)fluoranthene, benzo(k)fluoranthene, benzo(a)pyrene, indeno(1,2,3-  
303 cd)pyrene and benzo(g,h,i)perylene), whereas four were only found on filters from the diesel  
304 scenario (acenaphthylene, fluoranthene, chrysene and dibenz(a,h)anthracene). Among the PAHs in  
305 gas-phase (XAD-2), naphthalene was the only compound measured in all samples. Six other  
306 PAHs were detected in some of the samples, both from diesel and electric scenarios, with pyrene  
307 only detected in diesel scenario. PAHs with molecular weight higher than pyrene (MW=202) were  
308 not detectable in any of the XAD-2 tubes. Results in SI Table S5.

309

#### 310 *PM metal content*

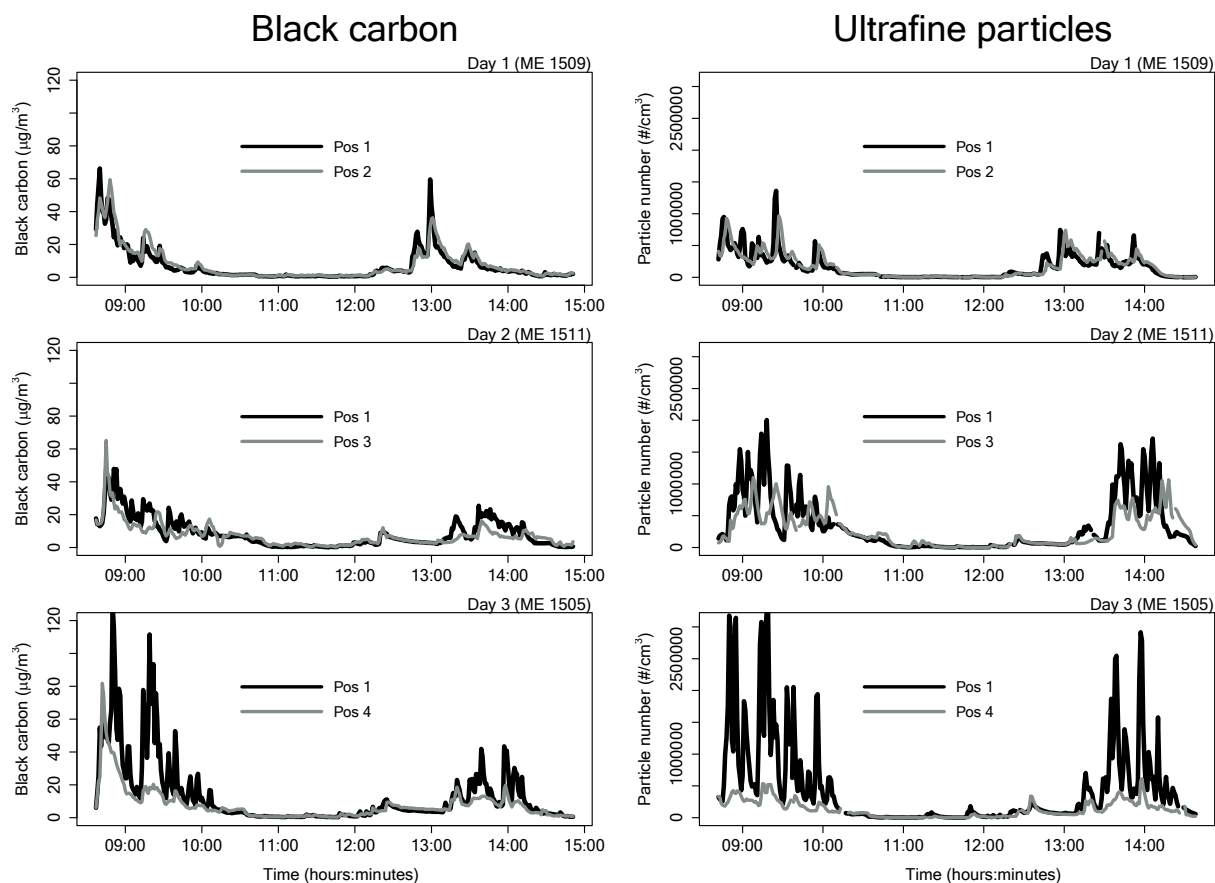


311 The results from elemental analysis of acid extracts from PM collected in diesel trains during 6  
312 days in December are presented in SI (Table S6) showing iron as the most abundant element  
313 measured (~45 mg/g), followed by zinc (~3 mg/g), magnesium (~2 mg/g) and copper (~1.5 mg/g),  
314 and with higher metal contents than the standard diesel exhaust particle SRM2975. The analysis  
315 was made using 2.35 mg PM of 100 mg collected in total.

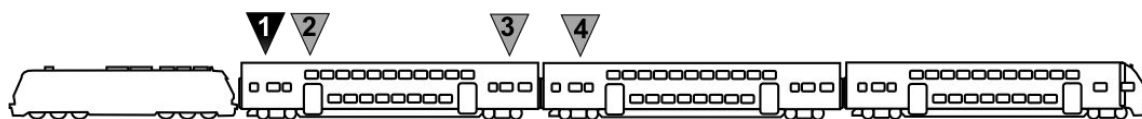
316

### 317 ***Gradient on-board the diesel train***

318 Figure 3 shows the concentrations of BC and UFP at four different positions inside the train.  
319 Position 1 was used as the reference position, where the volunteers were sitting in the main study  
320 (Figure S2). The measured concentrations at this first position and at position 2, just next to the  
321 first compartment, were similar, with a slightly lower average of the 6-hours of measurements on  
322 position 1, although the maximum values corresponded to position 1, both for particle number and  
323 BC mass concentrations (Tables S7 and S8). During the gradient measurements, an electrostatic  
324 sampler was placed at position 1 for the collection of PM. Even though the electrostatic sampler  
325 might have had an air cleaning effect at position 1, the peaked data reflected the proximity of  
326 position 1 to the source. Position 3 was in the end of the first train car and position 4 was in the  
327 beginning of the second car. The figure shows some day-to-day variation in concentrations, which  
328 is attributed to different engines, but overall it demonstrates a decrease in particle concentration  
329 with increasing distance from the engine.



330



331

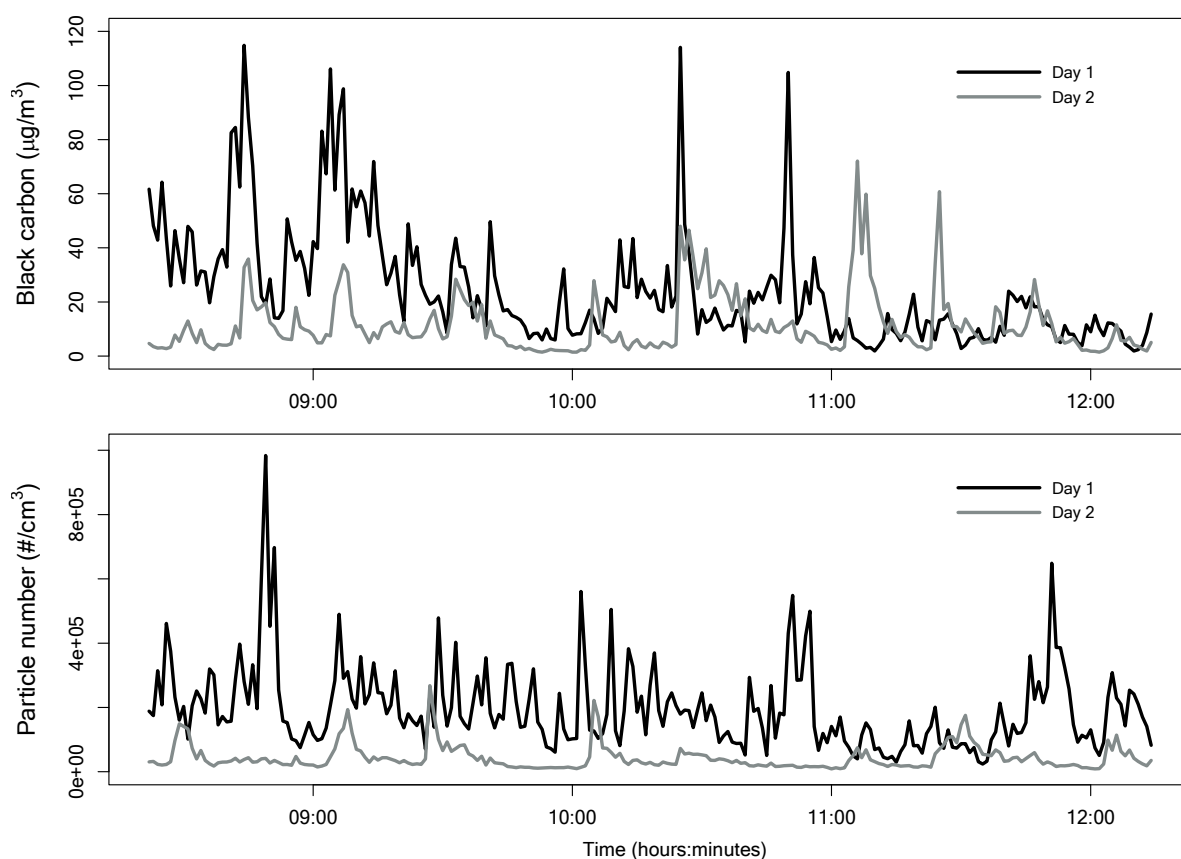
332 Figure 3 – Time-series data from BC and UFP concentrations (measured with DM) at four  
 333 positions inside the train during three days in December 2017 (Day 1, 13-12-2018; Day 2, 06-12-  
 334 2018; Day 3, 07-12-2018). Position 1 was in the front section of the first passenger car of the  
 335 diesel train, closest to the engine (Figure S2); position 2, was still in the first passenger car but  
 336 partially separated from position 1 with a glass that partially confines the front section (without  
 337 being closed); position 3, in the end of the first car; position 4, in the beginning of the second car.

338 Some short interruptions in the position other than position 1 correspond to device restarts  
 339 performed at stations.

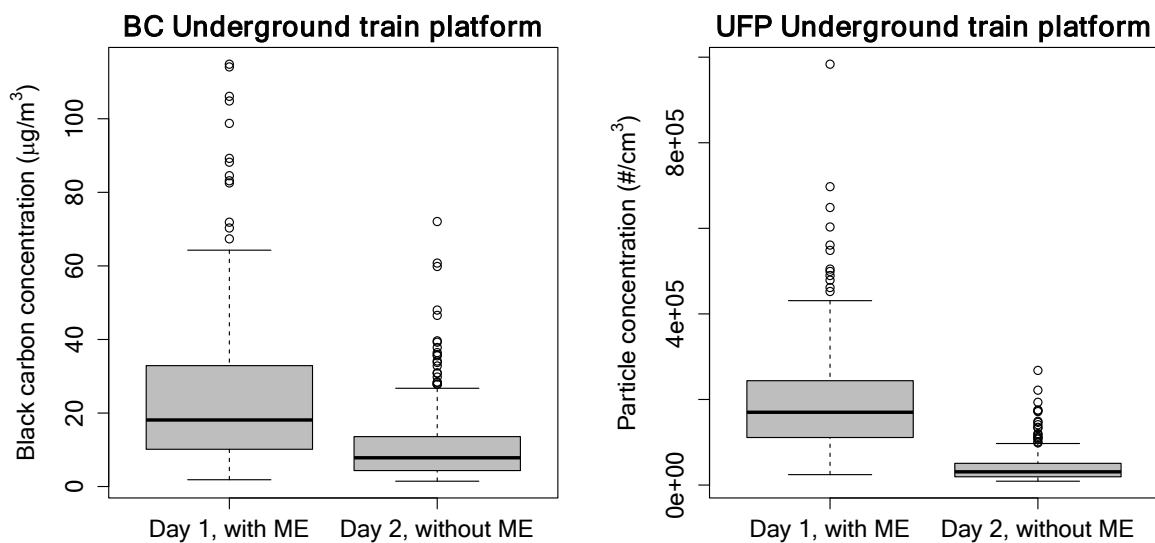
340

341 ***Underground station***

342 Figure 4 shows time-series of BC and UFP concentrations for two different days at Nørreport  
343 station underground train platform. The BC mass and UFP number concentrations were 2.2-fold  
344 and 4.5-fold higher, respectively, on the day the ME trains were in circulation (day 1) as compared  
345 to the day when the ME trains were taken out due to maintenance (day 2). Data for both days were  
346 recorded for the same time period and weekday, with slightly fewer trains in traffic in day 2  
347 (Figure 5). For day 2 the “other diesel” train passages were the ones that seemed to contribute the  
348 most to both mass and number concentrations (Table S9 and Figures S7 and S8).



349  
350 Figure 4- Time-series data from BC and UFP measured on underground train platform on two  
351 different Fridays, with and without ME trains in circulation. Day 1, with ME trains in circulation  
352 is represented in black colour and day 2, without the ME trains is represented in grey, both for  
353 time-series data with 1-minute resolution.



354

355 Figure 5- Underground train platform concentrations of black carbon mass (BC) and particle

356 number (UFP) on two different days with and without ME trains in circulation. Day 1

357 (26/01/2018) had 111 train passages during the 4 h of measurements (37 ME, 14 other diesel and

358 60 electric trains). Day 2 (02/02/2018) had 89 train passages during the 4 h of measurements (20

359 other diesel and 69 electric trains).

360

361 **Discussion**362 Concentrations of UFP, BC, NO<sub>x</sub>, NO<sub>2</sub>, PM<sub>2.5</sub> and BaP were higher in the passenger cars of diesel

363 trains as compared to electric trains. Considerably higher concentrations of UFP and BC were

364 measured when the locomotive was in pull mode. Our study presents a larger data set (more trips),

365 covers a longer time period and measures more components of DE than previous studies<sup>4, 7, 19</sup>. It is

366 also clear that older diesel trains contributed substantially to UFP and BC concentrations in an

367 underground station in Copenhagen.

368

369 The average increase in air concentrations of UFP inside the diesel train (around 200,000

370 particles/cm<sup>3</sup>) was considerably higher than the personal exposure measured during passive

371 transport (car, bus or train), average 17,000 particles/cm<sup>3</sup>, in a study of subjects in the general  
372 population in Copenhagen<sup>20</sup>. For a daily commuter in diesel trains it would give a significant  
373 contribution to the total exposure, depending on the time spent in transit. For train staff, the  
374 average excess BC exposure of 8 µg/m<sup>3</sup> would, if assumed it is equal to elemental carbon, for 8  
375 hours workday give rise to around 160 excess lung cancer deaths per 10,000 individuals over a  
376 lifetime, according to published exposure-response estimates<sup>12</sup>. Furthermore, environmental  
377 exposure to elemental carbon has been shown to increase the overall long-term mortality with  
378 around 6% per 1 µg/m<sup>3</sup> increase at the residence<sup>19</sup>.

379  
380 BaP is a potent carcinogen and air concentrations of BaP were 6-fold higher in diesel trains  
381 compared with electric trains (the same fold difference as for BC). The average concentrations of  
382 BaP in diesel trains (0.17 ng/m<sup>3</sup>) were higher than the reference level of 0.12 ng/m<sup>3</sup> (lifetime  
383 exposure) based on the World Health Organization (WHO) unit risk for lung cancer for PAH  
384 mixtures (additional lifetime cancer risk of 1 per 100,000)<sup>21</sup>. The average concentrations of PM<sub>2.5</sub>  
385 and NO<sub>2</sub> were 68 µg/m<sup>3</sup> and 53 µg/m<sup>3</sup>, respectively, inside the diesel trains. For a commuter who  
386 spends two hours per day travelling inside the first car of the diesel train, the average daily (24  
387 hour) personal exposure to PM<sub>2.5</sub> would increase from 15<sup>22</sup> to 19 µg/m<sup>3</sup> and for NO<sub>2</sub> from 20<sup>22</sup> to  
388 23 µg/m<sup>3</sup>. The WHO air quality guidelines for PM<sub>2.5</sub> is 25 µg/m<sup>3</sup> for 24 h and 40 µg/m<sup>3</sup> annual for  
389 NO<sub>2</sub>.

390  
391 The results from the real-time instruments (UFP and BC) showed that diesel exhaust enters the  
392 train passenger cars especially when the locomotive is in pull mode. It means that exposure to  
393 diesel exhaust in train passenger cars can be reduced considerably by using push rather than pull  
394 mode. Dramatic increases of UFP in pull compared with push mode have also been observed  
395 inside passenger cars of diesel trains in Israel (3 to 43 fold higher concentrations, measured in

396 different trains)<sup>4</sup> and in Canada (18 times higher concentrations, average across 28 trips)<sup>7</sup>. From  
397 the USA, the difference between pull and push mode was reported to be 3- and 9-fold in trains  
398 from New York and Boston, respectively<sup>19</sup>. Overall, the mean concentrations of UFP inside diesel  
399 trains in the present study were 20 and 35 times higher than in electric trains (measured with NT  
400 and DM, respectively), which is a larger difference than reported from previous studies<sup>6</sup>. Average  
401 concentrations of UFP and BC in first passenger car when locomotive was in pull mode in our  
402 study were about 450,000 particles/cm<sup>3</sup> and 20 µg/m<sup>3</sup>, respectively, which is higher than reported  
403 from Canada (UFP: 126,000 /cm<sup>3</sup> and BC: 17.8 µg/m<sup>3</sup>)<sup>7</sup>.

404  
405 Another observation from our study is the elevated BC and UFP concentrations in the first  
406 morning trip, demonstrated in the diesel A scenario time-series where the same locomotive was  
407 used all day (Figure 1). Higher emissions in the morning might be attributable to a “cold” engine  
408 having non-optimal combustion efficiency. The study also showed a decline in BC and UFP  
409 concentrations from the passenger car proximal to the locomotive to the more distal train cars,  
410 suggesting that infiltration of diesel exhaust in the first car is also spreading to other cars in the  
411 train, however being diluted along the way. The gradient measurements were performed in  
412 different days (different locomotives), which constitutes a limitation for the quantification and  
413 comparison of concentrations at each position. This exposure study was carried out in relatively  
414 old diesel trains from the 1980s and it may not be possible to extrapolate the observed results to  
415 other diesel trains.

416  
417 Cha and co-workers have reported a gradient in particle exposure inside electric railcars and noted  
418 the importance of interior ventilation for indoor air quality<sup>5</sup>. For the trains in our study, ventilation  
419 systems differ with season. In addition, there are two types of first passenger car in the double  
420 deck rolling stock, as well as two types of diesel fuel (winter and summer diesel). We did not have

421 enough information regarding ventilation to assess the impact on concentration levels, however  
422 during the 49 monitoring days on diesel scenarios we covered both ventilation systems, double  
423 deck cars and fuel type, and we always observed higher personal concentrations inside the  
424 passenger car for diesel compared to electric trains.

425

426 We observed a tendency of increased NO<sub>2</sub> concentrations in the end of the study, during winter, as  
427 might be expected<sup>23, 24</sup>, however we cannot conclude about any seasonal differences in relation to  
428 other parameters. The particle number concentration also appeared to increase in the colder  
429 months, but the contribution of the specific engine in use might be more important. As the trains  
430 move as far as 114 km from Copenhagen, local precipitation and wind information have also been  
431 difficult to investigate.

432

433 The aldehyde concentrations inside the trains were very low; however, the mean concentrations  
434 were slightly elevated in diesel compared with electric trains. The formaldehyde concentrations  
435 inside the trains were lower than the average personal exposure found among the general  
436 population in Sweden (19 µg/m<sup>3</sup>)<sup>25</sup>, and lower than average concentrations found in Danish  
437 dwellings (20 µg/m<sup>3</sup>)<sup>26</sup>.

438

439 The metal and metalloid elemental composition of the PM collected inside diesel trains showed  
440 enrichment especially in iron, with other major elements being zinc, magnesium and copper.

441 Although with higher concentrations, the pattern is similar to the standard reference material 2975,  
442 (Table S6) and also in comparison with other DE particulates<sup>27</sup>. The presence of silver was likely  
443 attributed to the electrostatic sampler silver-coating<sup>28</sup>.

444

445 The ME-locomotives will be replaced with new electric locomotives in 2021, according to a  
446 public announcement<sup>29</sup>, nevertheless abatement actions should be put on place until then.

447

#### 448 **Acknowledgments**

449 The technical assistance from Signe Nielsen, Ulla Tegner and Vivi Kofoed-Sørensen is gratefully  
450 acknowledged. Special thanks go to Danish State Railways for all the logistical support. We are  
451 also grateful to the volunteers carrying the monitoring devices for the considerable time and  
452 willingness put into this study.

453

#### 454 **Funding**

455 The research leading to these results has received funding from Danish Centre for Nanosafety II.

456

#### 457 **Supporting information**

458 Measurement details; detailed output of NanoScan SMPS; UFP measured with NT; detailed  
459 results from PAHs; detailed results from elemental analysis; and details of gradient and  
460 underground measurements. This information is available free of charge via the Internet at  
461 <http://pubs.acs.org>.

462

463

464

#### 465 **References**

- 466 1. OECD/IEA and UIC *Railway handbook 2017 - Energy consumption and CO2 emissions:*  
467 *focuss on passenger rail services*; [www.uic.org/uic-iea-railway-handbook](http://www.uic.org/uic-iea-railway-handbook), 2017; pp 1-  
468 120.
- 469 2. Burchill, M. J.; Gramotnev, D. K.; Gramotnev, G.; Davison, B. M.; Flegg, M. B., Monitoring  
470 and analysis of combustion aerosol emissions from fast moving diesel trains. *Sci Total*  
471 *Environ* **2011**, *409*, (5), 985-93.



- 472 3. Jaffe, D. A.; Hof, G.; Malashanka, S.; Putz, J.; Thayer, J.; Fry, J. L.; Ayres, B.; Pierce, J. R.,  
473 Diesel particulate matter emission factors and air quality implications from in-service  
474 rail in Washington State, USA. *Atmospheric Pollution Research* **2014**, *5*, (2), 344-351.
- 475 4. Abramesko, V.; Tartakovsky, L., Ultrafine particle air pollution inside diesel-propelled  
476 passenger trains. *Environ Pollut* **2017**, *226*, 288-296.
- 477 5. Cha, Y.; Tu, M.; Elmgren, M.; Silvergren, S.; Olofsson, U., Factors affecting the exposure of  
478 passengers, service staff and train drivers inside trains to airborne particles. *Environ Res*  
479 **2018**, *166*, 16-24.
- 480 6. Knibbs, L. D.; Cole-Hunter, T.; Morawska, L., A review of commuter exposure to ultrafine  
481 particles and its health effects. *Atmos Environ* **2011**, *45*, (16), 2611-2622.
- 482 7. Jeong, C. H.; Traub, A.; Evans, G. J., Exposure to ultrafine particles and black carbon in  
483 diesel-powered commuter trains. *Atmos Environ* **2017**, *155*, 46-52.
- 484 8. Abbasi, S.; Jansson, A.; Sellgren, U.; Olofsson, U., Particle Emissions From Rail Traffic: A  
485 Literature Review. *Crit Rev Env Sci Tec* **2013**, *43*, (23), 2511-2544.
- 486 9. Cha, Y. Y.; Abbasi, S.; Olofsson, U., Indoor and outdoor measurement of airborne  
487 particulates on a commuter train running partly in tunnels. *P I Mech Eng F-J Rai* **2018**,  
488 *232*, (1), 3-13.
- 489 10. IARC *Diesel and Gasoline Engine Exhausts and Some Nitroarenes.*; World Health  
490 Organization - International Agency for Research on Cancer:  
491 [www.ncbi.nlm.nih.gov/pubmed/26442290](http://www.ncbi.nlm.nih.gov/pubmed/26442290), 2014; pp 9-699.
- 492 11. Reis, H.; Reis, C.; Sharip, A.; Reis, W.; Zhao, Y.; Sinclair, R.; Beeson, L., Diesel exhaust  
493 exposure, its multi-system effects, and the effect of new technology diesel exhaust.  
494 *Environment International* **2018**, *114*, 252-265.
- 495 12. Vermeulen, R.; Silverman, D. T.; Garshick, E.; Vlaanderen, J.; Portengen, L.; Steenland, K.,  
496 Exposure-Response Estimates for Diesel Engine Exhaust and Lung Cancer Mortality  
497 Based on Data from Three Occupational Cohorts. *Environ Health Persp* **2014**, *122*, (2),  
498 172-177.
- 499 13. Karottki, G.; Loft, S. *Rapport vedrørende måling af udsættelse for ultrafine partikler blandt*  
500 *ansatte i DSB*; [www.dsb.dk/om-dsb/samfundsansvar/miljo/ultrafine-partikler/](http://www.dsb.dk/om-dsb/samfundsansvar/miljo/ultrafine-partikler/) on 08-  
501 10-2018, 2015; pp 1-48.
- 502 14. Madsen, A. M.; Sharma, A. K., Sampling of High Amounts of Bioaerosols Using a High-  
503 Volume Electrostatic Field Sampler. *The Annals of Occupational Hygiene* **2008**, *52*, (3),  
504 167-176.
- 505 15. Hagenbjork-Gustafsson, A.; Tornevi, A.; Forsberg, B.; Eriksson, K., Field validation of the  
506 Ogawa diffusive sampler for NO<sub>2</sub> and NO<sub>x</sub> in a cold climate. *J Environ Monit* **2010**, *12*,  
507 (6), 1315-24.
- 508 16. Kliucininkas, L.; Martuzevicius, D.; Krugly, E.; Prasauskas, T.; Kauneliene, V.; Molnar, P.;  
509 Strandberg, B., Indoor and outdoor concentrations of fine particles, particle-bound PAHs

- 510 and volatile organic compounds in Kaunas, Lithuania. *J Environ Monit* **2011**, *13*, (1), 182-  
511 91.
- 512 17. Andersen, M. H. G.; Saber, A. T.; Clausen, P. A.; Pedersen, J. E.; Løhr, M.; Kermanizadeh, A.;  
513 Loft, S.; Ebbenhøj, N. E.; Hansen, Å. M.; Pedersen, P. B.; Koponen, I. K.; Nørskov, E.-C.;  
514 Møller, P.; Vogel, U., Association between polycyclic aromatic hydrocarbons exposure  
515 and peripheral blood mononuclear cell DNA damage in human volunteers during fire  
516 extinction exercises. *Mutagenesis* **2018**, *33*, 105-115.
- 517 18. R Core Team, R: A language and environment for statistical computing. In R Foundation  
518 for Statistical Computing, Vienna, Austria: [www.R-project.org/](http://www.R-project.org/), 2014.
- 519 19. Clean Air Task Force *A multi-city investigation of exposure to diesel exhaust in multiple*  
520 *commuting modes.*; [www.catf.us/wp-](http://www.catf.us/wp-content/uploads/2019/02/CATF_Pub_Diesel_Exhaust_Exposure_Investigation.pdf)  
521 [content/uploads/2019/02/CATF\\_Pub\\_Diesel\\_Exhaust\\_Exposure\\_Investigation.pdf](http://www.catf.us/wp-content/uploads/2019/02/CATF_Pub_Diesel_Exhaust_Exposure_Investigation.pdf),  
522 2010; pp 1-84.
- 523 20. Beko, G.; Kjeldsen, B. U.; Olsen, Y.; Schipperijn, J.; Wierzbicka, A.; Karottki, D. G.; Toftum,  
524 J.; Loft, S.; Clausen, G., Contribution of various microenvironments to the daily personal  
525 exposure to ultrafine particles: Personal monitoring coupled with GPS tracking. *Atmos*  
526 *Environ* **2015**, *110*, 122-129.
- 527 21. WHO *WHO guidelines for indoor air quality: selected pollutants.*; World Health  
528 Organization: 2010.
- 529 22. Sorensen, M.; Loft, S.; Andersen, H. V.; Raaschou-Nielsen, O.; Skovgaard, L. T.; Knudsen, L.  
530 E.; Nielsen, I. V.; Hertel, O., Personal exposure to PM<sub>2.5</sub>, black smoke and NO<sub>2</sub> in  
531 Copenhagen: relationship to bedroom and outdoor concentrations covering seasonal  
532 variation. *J Expo Anal Environ Epidemiol* **2005**, *15*, (5), 413-22.
- 533 23. Grundström, M.; Linderholm, H. W.; Klingberg, J.; Pleijel, H., Urban NO<sub>2</sub> and NO pollution  
534 in relation to the North Atlantic Oscillation NAO. *Atmos Environ* **2011**, *45*, (4), 883-888.
- 535 24. Roberts-Semple, D.; Song, F.; Gao, Y., Seasonal characteristics of ambient nitrogen oxides  
536 and ground-level ozone in metropolitan northeastern New Jersey. *Atmospheric Pollution*  
537 *Research* **2012**, *3*, (2), 247-257.
- 538 25. Hagenbjork-Gustafsson, A.; Tornevi, A.; Andersson, E. M.; Johannesson, S.; Bellander, T.;  
539 Merritt, A. S.; Tinnerberg, H.; Westberg, F.; Forsberg, B.; Sallsten, G., Determinants of  
540 personal exposure to some carcinogenic substances and nitrogen dioxide among the  
541 general population in five Swedish cities. *J Expo Sci Env Epid* **2014**, *24*, (4), 437-443.
- 542 26. Raaschou-Nielsen, O.; Hermansen, M. N.; Loland, L.; Buchvald, F.; Pipper, C. B.; Sorensen,  
543 M.; Loft, S.; Bisgaard, H., Long-term exposure to indoor air pollution and wheezing  
544 symptoms in infants. *Indoor Air* **2010**, *20*, (2), 159-167.
- 545 27. Loxham, M.; Cooper, M. J.; Gerlofs-Nijland, M. E.; Cassee, F. R.; Davies, D. E.; Palmer, M. R.;  
546 Teagle, D. A. H., Physicochemical Characterization of Airborne Particulate Matter at a  
547 Mainline Underground Railway Station. *Environ Sci Technol* **2013**, *47*, (8), 3614-3622.

548 28. Sharma, A. K.; Wallin, H.; Jensen, K. A., High volume electrostatic field-sampler for  
549 collection of fine particle bulk samples. *Atmos Environ* **2007**, *41*, (2), 369-381.

550 29. DSB *Fremtidens tog - beslutningsoplæg for Fase 1.5*;  
551 [https://www.dsb.dk/globalassets/om-dsb/rapporter/fremtidens-  
tog/beslutningsoplag-for-fase-1.5.pdf](https://www.dsb.dk/globalassets/om-dsb/rapporter/fremtidens-<br/>552 tog/beslutningsoplag-for-fase-1.5.pdf) on 06-01-2019, 2017.

553