Isoprene photo-oxidation products quantify the effect of pollution on hydroxyl radicals over Amazonia

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Published in:
Science Advances

DOI:
10.1126/sciadv.aar2547

Publication date:
2018

Document version
Publisher's PDF, also known as Version of record

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Citation for published version (APA):
Nitrogen oxides (NO$_x$) emitted from human activities are believed to regulate the atmospheric oxidation capacity of the troposphere. However, observational evidence is limited for the low-to-median NO$_x$ concentrations prevalent outside of polluted regions. Directly measuring oxidation capacity, represented primarily by hydroxyl radicals (OH), is challenging, and the span in NO$_x$ concentrations at a single observation site is often not wide. Concentrations of isoprene and its photo-oxidation products were used to infer the equivalent noontime OH concentrations. The fetch at an observation site in central Amazonia experienced varied contributions from background regional air, urban pollution, and biomass burning. The afternoon concentrations of reactive nitrogen oxides (NO$_x$), indicative of NO$_x$ exposure during the preceding few hours, spanned from 0.3 to 3.5 parts per billion. Accompanying the increase of NO$_x$ concentration, the inferred equivalent noontime OH concentrations increased by at least 250% from $0.6 \times 10^6$ to $1.6 \times 10^6$ cm$^{-3}$. The conclusion is that, compared to background conditions of low NO$_x$ concentrations over the Amazon forest, pollution increased NO$_x$ concentrations and amplified OH concentrations, indicating the susceptibility of the atmospheric oxidation capacity over the forest to anthropogenic influence and reinforcing the important role of NO$_x$ in sustaining OH concentrations.

**INTRODUCTION**

Earth’s atmosphere is an oxidizing medium that drives organic molecules toward carbon dioxide, and oxidation by hydroxyl radicals (OH) initiates most of these reactions (1, 2). Oxidation by OH radicals also leads to the production of many secondary pollutants that affect human health and climate, such as organic particulate matter and ozone. In relation to OH concentrations, the NO$_x$ family, defined as including nitric oxide (NO) and nitrogen dioxide (NO$_2$), has two roles (2). The chemistry is illustrated in Fig. 1A. On the one hand, NO reacts with hydroperoxyl radicals (HO$_2$) and organic peroxy radicals (RO$_2$) to produce OH catalytically in the presence of sunlight, thereby enhancing OH concentrations. On the other hand, at higher NO$_x$ concentrations, the direct reaction of NO$_x$ with OH to produce nitric acid (HNO$_3$) becomes dominant, and NO$_x$ serves as an OH sink. Hydroxyl radical concentrations thus increase as NO$_x$ concentrations increase below a threshold NO$_x$ concentration, and they decrease as NO$_x$ concentrations increase above this threshold. The net result is that OH concentrations follow a bell curve with respect to NO$_x$ concentrations, as illustrated in Fig. 1B (2, 3).

The applicability of this classical understanding of OH-NO$_x$ chemistry, as represented by the bell curve, to atmospheric conditions is under challenge based on field measurements of OH concentrations in regions where volatile organic compounds (VOCs) are abundant. A meta-study examined the dependence of OH concentrations across a broad range of NO$_x$ conditions (4). The observations ranged from tropical forests in South America and Southeast Asia (5, 6) to deciduous forest in the United States and rural area in China (7–9) and to polluted metropolitan regions of New York City, Beijing, Tokyo, and Mexico.
City (10–13). Above a threshold NOx concentration, OH concentrations decreased with increasing NOx concentration, as expected. However, below the threshold NOx concentration, reported OH concentrations in many regions were unexpectedly high, and collectively, they appeared to be independent of NOx concentration, as illustrated by the horizontal line in Fig. 1B.

Possible mechanisms for maintaining elevated OH concentrations under atmospheric conditions with below-threshold NOx concentrations have been considered (7, 14–16). In regions where VOCs are abundant, especially over and downwind of forests, the major fate of OH is reaction with biogenic VOCs to produce RO2. Possible reactions of RO2 radicals to regenerate OH and thereby maintain OH concentrations have been suggested, especially for forested regions dominated by isoprene emissions, such as OH production from the reaction of HO2 with isoprene-derived RO2 radicals (ISOPOO) (14, 17), as well as OH release following ISOPOO isomerization (16, 18). Even so, these mechanisms appear insufficient to quantitatively explain the reported high OH concentrations at low NOx concentrations (14, 17, 18). Alternatively, the accuracy of reports of high OH concentrations is uncertain because of the possibility of interferences in the underlying measurement technique of laser-induced fluorescence (LIF) (9, 19, 20). For instance, Mao et al. (9) attributed 40 to 60% of the nominal OH signal observed in a California forest to interferences. After correction, the data fell on the bell curve, instead of on the horizontal line of Fig. 1B (4). To date, there is no consensus on whether similar interferences may have occurred in the other measurements included in the meta-study. The LIF instruments are custom-built in variable configurations by different research groups. Another uncertainty related to the meta-study is that the span in NOx concentration of each underlying study was usually not wide, so no individual study closely tested the response of OH concentration to NOx concentration across a broad range. Instead, the analysis in the meta-study was based on normalization of observations across multiple sites, different instruments, and differing atmospheric conditions. In short, possible new mechanisms for OH production and recycling under atmospheric conditions and possible artifacts in the underlying data sets all remain to be reconciled.

Herein, a complementary approach based on isoprene photo-oxidation products is presented for mapping the dependence of OH concentration on NOx concentration for atmospheric conditions over an isoprene-dominated forested environment. Precedent approaches for estimating OH concentrations include the use of a range of OH-reacting trace species, as well as OH reaction products and their ratios (3, 21–24). Isoprene is the dominant VOC emitted to the atmosphere from many forests (25), and reaction with OH is its primary loss pathway (26). As OH concentration increases, the concentration CISOP of isoprene decreases, and the sum concentration CPROD of its oxidation products increases, provided that other factors, such as reaction time, ozone concentration, and isoprene emission rates, are unchanged. The presentation herein develops an analysis to use the afternoon parent-to-product concentration ratio CISOP/CPROD to infer equivalent noontime OH concentration within an air mass during the preceding daylight hours. The analysis must account for several factors, in addition to OH concentration, that influence CISOP/CPROD.

Data sets were recorded at a single site in the Observations and Modeling of the Green Ocean Amazon (GoAmazon2014/5) Experiment in central Amazonia during the wet and dry seasons of 2014 (27). At different times, the sampling site received mixtures to various degrees of unpolluted regional background air, air influenced by regional fires (especially in the dry season), or air that had passed over the nearby city of Manaus, an urban region of 2 million inhabitants and a strong regional source of NOx. The site was 4 to 6 hours downwind from Manaus for the typical prevailing easterlies associated with trade winds (27, 28). As a result, the GoAmazon2014/5 data sets spanned a great breadth of upwind NOx chemistry at a single observation site (29). The range in measured isoprene concentrations was similar to that of the forested sites of the meta-study (4). The GoAmazon2014/5 data sets were used in the analysis herein to strongly test for a bell curve versus a horizontal line in the response of OH concentration to NOx concentration in an isoprene-dominated forested environment (Fig. 1B).

**RESULTS**

The median and interquartile ranges of measured CISOP and CPROD are plotted by hour in Fig. 2 (A and B). The dry and wet season data sets are represented by red and blue colors, respectively. The product concentration CPROD represents the sum concentration of methyl vinyl ketone (MVK, C4H6O), methacrolein (MACR; C4H6O), and isoprene-derived hydroperoxide isomers [including (1,2)-ISOPOOH and (4,3)-ISOPOOH, hereafter ISOPOOH; C4H8O5]. MVK and MACR are the major OH-reaction products of isoprene along NOx-mediated pathways (17).
ISOPOOH species are the major products along HO2-mediated pathways (30). Although the relative importance of the NO and HO2 pathways varies with NOx concentration (29), the summed production yield of MVK, MACR, and ISOPOOH is approximately 70% across the atmospherically relevant range of NOx concentrations (30).

Figure 2 (A and B) shows that before sunrise, CISOPEP and CPROD were <20 and <10% of their respective afternoon values. At sunrise, the forest began to emit isoprene, and photochemical production of OH also began. Isoprene had a lifetime of 2.8 hours against OH attack for \( C_{OH} = 1.0 \times 10^6 \text{ cm}^{-3} \), whereas the lifetimes of products against chemical loss were up to five times higher. In line with these processes, Fig. 2 (A and B) shows that at sunrise, CISOPEP increased, followed shortly thereafter by increases in CPROD. Isoprene also reacted with photochemically produced O3, but this pathway was typically one order of magnitude less important than the OH pathway for the prevailing atmospheric conditions (26). The resulting MVK and MACR was a minor contribution to the overall production of these species. In addition to chemical loss and production, CISOPEP and CPROD were also affected by dry and wet deposition, as well as by entrainment of air from above the boundary layer. The analysis herein focuses on afternoon concentrations CISOPEP and CPROD because OH oxidation has the greatest impact on CISOPEP and CPROD at this time of the day. In addition, the possibility of intercepting a polluted air mass from Manaus was highest in the afternoon, given that the research site was 4 to 6 hours downwind of this pollution source for typical winds following sunrise (28).

Scatterplots for the afternoon hours (13:00 to 16:00 local time) of CISOPEP, CPROD, and CPROD/CISOPEP in relation to the sum concentration CNOx of reactive nitrogen species (NOx) are shown in Fig. 3, A to C, respectively. In addition to the NOx species, the NOx family also included reservoir forms of oxidized nitrogen, such as HNO3, which were products of atmospheric reactions of NOx (3). A few hours downwind of the Manaus source region at the observation site, CNOx remained a semiconserved quantity, whereas CISOPEP had significantly decreased. Atmospheric conversion of NOx to NOy also occurred for air masses downwind of regional biomass burning. Therefore, for the analysis, CISOPEP was used to represent the integrated intensity of NOx chemistry before arrival of an air mass at the observation site (29, 31). The ratio CISOPEP/CNOx had an interquartile variability of 0.30 to 0.44 across the study period, supporting the use of the simplified representation CISOPEP. As shown in Fig. 3, the hourly averaged afternoon values of CNOx spanned more than one order of magnitude from 0.3 to 3.5 parts per billion (ppb). The interquartile range of CNOx in the afternoon hours was 0.7 to 1.2 ppb in the wet season and 1.3 to 2.0 ppb in the dry season. As a reference point, CNOx for regional background conditions was 0.46 ± 0.26 ppb (32). Concentrations above 1 ppb at the observation site indicated the effects of Manaus pollution or regional biomass burning (29, 31).

The scatterplots of Fig. 3 illustrate the following findings. CISOPEP did not correlate with CNOx (Spearman’s rank \( r = 0.07 \); Fig. 3A), CPROD did correlate with CNOx (\( P < 10^{-4} \); Fig. 3B), yet the data were scattered (\( r = 0.4 \)). By comparison, the ratio CPROD/CISOPEP correlated tightly with CNOx (\( P < 10^{-4} \) and \( r = 0.6 \); Fig. 3C). The median of CPROD/CISOPEP increased from 0.4 to 1.0 as CNOx changed from below 0.5 to above 2 ppb. A high value of CPROD/CISOPEP was never observed for a low value of CNOx (Fig. 3C).

An unclear trend in the concentration plots on the one hand (Fig. 3, A and B) compared to a clear trend in the ratio plot on the other hand (Fig. 3C) with respect to CNOx can be explained by the large temporal variability in isoprene emissions (22). This variability was independent of CNOx and, thus, confounded direct relationships between CISOPEP and CNOx, or CPROD and CNOx. By comparison, the ratio CPROD/CISOPEP largely compensated the variability in isoprene emissions and thereby revealed differences in atmospheric oxidation. For these reasons, CPROD/CISOPEP was used in the further analysis herein, focusing on understanding and quantifying the effects of pollution on the atmospheric oxidation cycle over central Amazonia.

**DISCUSSION**

Relating OH concentration to the ratio CPROD/CISOPEP

A model of isoprene photochemistry over the course of 1 day was constrained by the observed values of CPROD/CISOPEP to estimate the equivalent noontime OH concentration for the preceding daylight.
hours upwind of the observation site, as follows. For products species \(i\) of isoprene oxidation, where \(i\) is one of (1,2)-ISOPOOH, (4,3)-ISOPOOH, MVK, or MACR, the time course of product concentrations \(C_i(t)\) in an air mass is governed by the following family of equations

\[
dC_i/dt = y_i(NO) k_{ISOPOOH} C_{OH}(t) C_{ISOPOH}(t) + y_{i,O} k_{ISOPOOH} C_{O}(t) C_{ISOPOH}(t) - k_i(t) C_i(t)
\]

where \(y_i\) is the production yield of product species \(i\) from the reaction between hydroxyl radical and isoprene. It varies according to the fate of ISOPOOH, which is largely controlled by the NO concentration and hence susceptible to pollution. \(C_{OH}(t), C_{ISOPOH}(t),\) and \(C_{O}(t)\), where \(t\) represents time, are the concentrations of hydroxyl radical, isoprene, and ozone, respectively, all of which vary strongly with time of day. \(k_{ISOPOOH}\) and \(k_{ISOPOH}\) are the reaction rate constants of isoprene with hydroxyl radical and ozone, respectively. \(y_{i,O}\) is the production yield of product species \(i\) from the reaction between ozone and isoprene \((30)\). \(k_i\) is a composite, pseudo-first-order loss coefficient of species \(i\), given by \(k_i(t) = k_{i,OH} C_{OH}(t) + k_{i,O} C_{O}(t) + k_{i,H} + k_{i,de}\) for bimolecular reaction between species \(i\) and OH or O\(_3\), entrainment mixing with air above the boundary layer (\(en\)), and deposition to the planetary surface (\(de\)).

Table S1 lists values of \(y_{i,O}, k_{i,OH}, k_{i,O}\), as well as typical values of \(k_{i,H}\) and \(k_{i,de}\) for central Amazonia, partly reproduced from Liu et al. \((29)\). For comparison, a more detailed model for entrainment to couple chemistry and boundary layer dynamics was also explored, and similar results were obtained (see the Supplementary Materials).

The text below first discusses the sequential reaction model (SRM), which is a commonly used simplification of Eq. 1 to relate the observed \(C_{ISOPOH}\) to an inferred \(OH\) concentration \((21)\). The conclusion reached is that the SRM is not appropriate for analysis of the collected data set. The text then introduces a complementary approach based on (i) time-dependent approximations of \(C_{ISOPOH}, C_{OH},\) and \(C_{O}\), and (ii) substitution of the production yield \(y_i\) by an effective production yield denoted by \(y'_i\) \((C_{SO})\).

Analysis by SRM assumes that isoprene emissions can be taken as a point source (typically an upwind forest) and that isoprene is subsequently oxidized during transport to an observation site. More specifically, isoprene is not emitted into the air parcel along its path of travel, and reaction with OH is the dominant process governing the time evolution of \(C_{ISOPOH}\) (that is, \(dC_{ISOPOH}/dt = -k_{ISOPOOH} C_{OH} C_{ISOPOH}\)). Other assumptions include constant \(C_{OH}\) and \(C_{O}\), a fixed \(y_i\), a fixed reaction time \(t\), and negligible entrainment and deposition. For these conditions, Eq. 1 transforms to an explicit relation that allows \(C_{OH}\) to be inferred from measured \(C_{ISOPOH}/C_{ISOPOH}\) \((21)\). The major underlying SRM assumption of an upwind point source was, however, not applicable to the GoAmazon2014/5 scenario. The observation site was surrounded by forest for hundreds of kilometers, meaning that isoprene was continuously emitted into air parcels throughout transport and that \(C_{ISOPOH}\) and \(C_{ISOPOH}\) corresponded to the integrated balance between source and loss processes throughout transport.

An alternative scheme to the SRM is developed herein to constrain \(C_{OH}\) based on observed \(C_{ISOPOH}/C_{ISOPOH}\). The regional area around T3 is approximated as a homogeneous forest representing a perfectly diffuse nonpoint source region of isoprene. Analysis using Google Earth shows that surface forest coverage exceeded 70% in a 100-km radius for the dominating easterlies and northeasterlies in the wind rose and >80%

for all directions \((33)\). The radius of 100 km was based on typical wind speeds of 10 to 20 km h\(^{-1}\) and the time period from the start of photochemistry at daybreak to the analysis window in the afternoon. Within this 100-km radius, the forest type and, hence, isoprene emissions varied to some extent \((34)\), and there were also scattered pastures, two large rivers, and the urban area of Manaus. These factors notwithstanding, a homogeneous diffuse source of isoprene emissions was taken as an acceptable approximation for the accuracy of the modeling herein.

Under this approach of a homogeneous source region, a time series of observations at a Eulerian point is fully transformable into a Lagrangian model of a time series of concentrations within an air parcel over the course of a time period \((35, 36)\). Equation 1 describes the transformations of product species \(C_i\) within the Lagrangian parcel, and in the treatment herein, observations constrained the terms. For instance, measurements of the time course of isoprene concentrations directly constrained \(C_{ISOPOH}(t)\). Steady increases were observed from sunrise to midafternoon (Fig. 2A), and the observed time dependence was represented empirically by the following linear equation

\[
C_{ISOPOH}(t) = \xi C_{ISOPOH,0} \left(1 + t/t_{ISOPOH}^\prime\right)
\]

where \(\xi\) was a daily scaling factor representing the variability in isoprene concentrations for each day, \(C_{ISOPOH,0}\) was the typical isoprene concentration at time zero (sunrise), and \(t_{ISOPOH}^\prime\) was the typical characteristic time for \(C_{ISOPOH}\) to double its initial value. A value of 2.5 hours for \(t_{ISOPOH}^\prime\) represented most days. Values of \(C_{ISOPOH,0}\) of 0.35 and 0.6 ppb were used as approximate values to represent the wet and dry seasons, respectively. Equation 2 is plotted in Fig. 2A as the black dashed lines to represent median \(C_{ISOPOH}(t)\) from sunrise to midafternoon in the two seasons using \(\xi\) of unity. The analysis below further shows that the quantities \(\xi\) and \(C_{ISOPOH,0}\) drop out (Eq. 5) so that the actual values are not important.

In regard to \(C_{OH}(t),\) hydroxyl radical concentrations strongly correlate with the photolysis frequency \(J_O3,\) of ozone for most locations worldwide \((37)\). In the analysis, the daily course of hydroxyl radical concentrations was approximated as an equivalent noontime concentration \(C_{OH,noon}\) modulated by a time-dependent photolysis frequency \(J_O3,\) of ozone along the O\((1D)\) channel, as follows

\[
C_{OH}(t) = C_{OH,noon} J_O3(t) / J_O3,noon
\]

where \(J_O3,noon\) was the peak value occurring at local noon for the equatorial location. The variation of \(J_O3(t)/J_O3,noon\) followed that of the Master Chemical Mechanism for clear skies at the latitude and longitude of the observation site (fig. S2) \((30)\). The value of \(C_{OH,noon}\) was not known before the analysis, and inferring its value was a main point of the analysis presented herein.

The production yield \(y_i\) \((C_{NO}(t))\) in Eq. 1, written in full form as \(y_i[C_{NO}(t)]\), does not fulfill the homogeneity requirement of a Eulerian- to-Lagrangian transformation because of the decreasing concentration of NO along the path. The observations and analysis presented in the study of Liu et al. \((29)\) for this same observation site showed that the ratio of product concentrations \(C_{ISOPOOH}/C_{MVK+MACR}\) correlated tightly with \(C_{NO}\), and this ratio was further transformed in that study to the effective production ratio \(y_{ISOPOOH,y_{MVK+MACR}}\) by combining the measurements with kinetic modeling \((29)\). Herein, the production...
yield \( y_i[C_{NO}(t)] \) in the Lagrangian framework was substituted by an empirical effective production yield \( y_i^*(C_{NO}) \). The relationship \( y_i^*(C_{NO}) \) is presented in fig. S3 (see further in the Supplementary Materials). This treatment subsumed the reality of a detailed history of NO exposure within the sampled air parcel along its Eulerian path into an effective behavior, which was empirically quantified by semi-conserved \( C_{NO} \) during the course of Lagrangian time. Although this approach was approximate, the tight relationship of concentration ratio and \( C_{NO} \) observed in the study of Liu et al. (29) supported its use for the analysis herein.

The quantity \( C_{O_3}(t) \) in Eq. 1 was directly constrained by two observations: (i) the increase of ozone concentrations from sunrise to midafternoon (fig. S4A) and (ii) the correlation between \( C_{O_3} \) and \( C_{NO} \) for afternoon time periods (fig. S4B). On the basis of these observations, \( C_{O_3}(t) \) for an air parcel arriving at the observation site in the afternoon was empirically approximated as a function of observed concentration factor \( y \) (fig. S4B). On the basis of these observations, \( C_{O_3}(t) \) was represented by \( C_{O_3}^*(C_{NO},t) \), as explained further in the Supplementary Materials. Overall, sensitivity tests presented herein show that the influence of ozone chemistry on \( C_{PROD}/C_{ISOP} \) and, hence, inferred \( C_{OH,noon} \) was small.

For these treatments of \( C_{ISOP}(t) \), \( C_{OH}(t) \), \( C_{O_3}(t) \), and \( y_i(\text{NO}) \), Eq. 1 can be rewritten as follows:

\[
\frac{dC_i}{dt} = \left[ y_i^*(C_{NO}) \right] k_{ISOP,OH} C_{OH,noon} \quad \frac{f_{O_3}(t)}{f_{O_3,noon}} + \frac{1}{k_{ISOP} - \xi} \left( 1 + \frac{t}{t_{ISOP}} \right) - \left[ k(t) \right] C_i(t) \quad y_i(0) = 0
\]

The initial concentrations of product species \( i \) are taken as zero [that is, \( C_i(0) = 0 \)] based on the data sets of Fig. 2. The symbol \( \xi \), defined by \( \xi = C_i/C_{ISOP} \), is introduced for the product ratio. By definition, \( C_{PROD}/C_{ISOP} = \Sigma_i \xi_i \). The following equation holds for \( d\xi_i/dt \):

\[
\frac{d\xi_i}{dt} = y_i^*(C_{NO}) \frac{k_{ISOP,OH} C_{OH,noon} f_{O_3}(t)}{f_{O_3,noon}} + y_i(1) \frac{k(t)}{k_{ISOP} + \xi(t)} - \left[ k(t) \right] \xi_i(t) \quad \xi_i(0) = 0
\]

Equation 5 is derived in the Supplementary Materials. The scaling factor \( \xi \) and the initial concentration \( C_{ISOP} \) drop out. Equation 5 captures the behavior shown in Fig. 3 that the ratio \( C_{PROD}/C_{ISOP} \) is independent of isoprene emissions even as the product concentration \( C_{PROD} \) is not. Equation 5 also suggests that the NO exposure of the air mass, represented by \( C_{NO} \), here, is a key driver of \( C_{PROD}/C_{ISOP} \) under ambient conditions through (i) effects on \( y_i^*(C_{NO}) \), (ii) effects on \( C_{O_3}(C_{NO}) \), and (iii) possible effects on \( C_{OH,noon}(C_{NO}) \).

The relative sensitivity of the analysis to \( y_i^*(C_{NO}) \), \( C_{O_3}(C_{NO}) \), and \( C_{OH,noon}(C_{NO}) \) is examined in Fig. 4. The figure shows \( C_{PROD}/C_{ISOP} \) as a function of \( C_{NO} \) for two fixed values of \( C_{OH,noon} \), specifically \( 1 \times 10^6 \) and \( 2 \times 10^6 \) cm\(^{-3} \), using Eq. 5 with and without ozone chemistry. At fixed \( C_{OH,noon} \), \( C_{NO} \) affects \( C_{PROD}/C_{ISOP} \) via the effects on \( y_i^*(C_{NO}) \) and \( C_{O_3}(C_{NO}) \) (Eq. 5). Solid lines show results for the full model of Eq. 5, and dashed lines show results for a model that omits isoprene ozonolysis (that is, \( C_{O_3} = 0 \)). For all cases, \( t = 8.5 \) hours, corresponding to 14:30 (local time).

Fig. 4. Simulated dependence of the concentration ratio \( C_{PROD}/C_{ISOP} \) on \( C_{NO} \). Results are shown for two different values of \( C_{OH,noon} \), as well as the full model compared to a model that omitted isoprene ozonolysis. At fixed \( C_{OH,noon} \), \( C_{NO} \) affects \( C_{PROD}/C_{ISOP} \) via the effects on \( y_i^*(C_{NO}) \) and \( C_{O_3}(C_{NO}) \) (Eq. 5). Solid lines show results for the full model of Eq. 5, and dashed lines show results for a model that omits isoprene ozonolysis (that is, \( C_{O_3} = 0 \)). For all cases, \( t = 8.5 \) hours, corresponding to 14:30 (local time).
Entrainment of air aloft accompanying the growth of the convective boundary layer is another important process that can affect the inferred $C_{\text{OH,noon}}$ values. Sensitivity tests were performed to evaluate the treatment embedded in $k(t)$ of Eq. 1 as compared to a more detailed mixing model. The results, presented in the Supplementary Materials, suggest that $C_{\text{OH,noon}}$ can be underestimated on the order of 20% for higher values of $C_{\text{NO}_y}$, again indicating that the line drawn in Fig. 5 can be an underestimate.

Random measurement errors of $C_{\text{PROD}}$, $C_{\text{ISO}}$, and $C_{\text{NO}_y}$, and systematic errors in the parameterizations of $C_{\text{OH}}(t)$, $C_{\text{ISO}}(t)$, and $y'(C_{\text{NO}_y})$ were also considered. The methods are presented in the Supplementary Materials (see table S2). As a result, inferred $C_{\text{OH,noon}}$ values are estimated to have a standard error of ±30%. This uncertainty is small relative to the trend of an increase by 250% of $C_{\text{OH,noon}}$ from small to high $C_{\text{NO}_y}$.

In the context of the controversy involving direct OH observations (9, 19, 20), the inference of OH concentrations from measurements of isoprene and its oxidation products, as presented herein, represents an important, albeit indirect, complementary analysis. The result presented in Fig. 5 serves as counterevidence to the conclusion of the earlier meta-study, suggesting that OH concentrations are independent of $C_{\text{NO}_y}$ in low-$C_{\text{NO}_y}$, high-isoprene environments (4). The OH concentrations obtained herein were 80% lower than those reported using the LIF technique over rainforest in coastal eastern South America for similar isoprene concentrations. By comparison, the OH concentrations inferred by the analysis herein were comparable to concurrent GoAmazon2014/5 observations using chemical ionization mass spectrometry deployed at the same observation site (see the Supplementary Materials). In addition, the OH concentrations fell into the range of OH concentrations inferred from previous airborne measurements over tropical forests in South America (see table S3 and references therein).

In summary, the current study shows that OH concentrations inferred from measurements of isoprene and its oxidation products increased with increasing $C_{\text{NO}_y}$ concentrations over a tropical forest in the central Amazon basin, in support of the bell curve response presented in Fig. 1. This result suggests that the oxidation capacity over tropical forests is susceptible to anthropogenic $C_{\text{NO}_y}$ emissions. Looking into the future in Amazonia, trends of ongoing deforestation for agricultural purposes and expanding urbanization can be expected to increase demand for electricity and transport (39). Increased $C_{\text{NO}_y}$ emissions should be expected on the basis of current technologies. Hydroxyl radical concentrations can be expected to likewise increase, keeping other factors equal, in forested regions affected by anthropogenic pollution in Amazonia. Increased OH concentrations imply changed spatial and temporal oxidation patterns for VOCs emitted from the forest, which can have follow-on effects on visibility, cloud formation, and rainfall based on shifted mass concentration, size distribution, and chemical composition of organic particulate matter (28, 40).

**MATERIALS AND METHODS**

Measurements were made at the “T3” site of the GoAmazon2014/5 Experiment (27). The T3 site was located in a pasture area of 2.5 km × 2 km in central Amazonia (−3.2133°, −60.5987°). Pasture regions have low emissions of isoprene (25). The site was 70 km west of Manaus, Brazil. Steady equatorial trade winds passed over Manaus in the direction of T3. The region between the city and T3 was largely forested with interspersed agricultural activities, and the isoprene observed at T3 originated from the upwind forest. Depending on variability in the winds,
the fetch of air at T3 varied from background conditions of the tropical forest (that is, low NO$_3^-$ to polluted conditions under the influence of Manaus and regional biomass burning (that is, higher NO$_3^-$)). Measurements were made during two intensive operating periods that took place in the wet season (01 February 2014 to 31 March 2014; IOP1) and in the dry season (15 August 2014 to 15 October 2014; IOP2).

A proton-transfer-reaction time-of-flight mass spectrometer (PTR-TOF-MS; Ionicon Analytik GmbH) was used to measure isoprene and its oxidation products (29). The PTR-TOF-MS was part of the Mobile Aerosol Observing System (MAOS) of the U.S. Department of Energy (DOE) (27). Isoprene (C$_{5}$H$_{10}$) was detected as the C$_{4}$H$_{7}^{+}$ ion. MVK (C$_{4}$H$_{6}$O), MACR (C$_{4}$H$_{6}$O), and two major isoprene-derived hydroperoxide isomers, (1,2)-ISOPOOH and (4,3)-ISOPOOH (C$_{3}$H$_{4}$O$_{3}$), were collectively detected as the C$_{3}$H$_{4}$O$_{2}^{-}$ ion (29). The instrument response was calibrated using authentic standards of these compounds. The product analysis herein was based on the sum concentration of MVK, MACR, and the two ISOPOOH isomers.

Instrumentation for measuring concentrations of reactive nitrogen oxides (NO$_x$) was part of MAOS. The NO$_x$ data sets were obtained from the Atmospheric Radiation Measurement (ARM) data archive (27). The operational detection limit of NO$_2$ was 0.1 ppb. The NO$_2$ concentrations were smoothed by applying a 30-min median filter to minimize the contribution of any local transient emissions, such as vehicles. Other complementary measurements at the site, such as meteorological parameters, were also obtained through the ARM data archive (27).

SUPPLEMENTARY MATERIALS

Supplementary material for this article is available at http://advances.sciencemag.org/cgi/content/full/4/4/eaar2547/DC1

section S1. Use of observed C$_{5}$H$_{9}O$ to represent the average NO$_x$ exposure

section S2. Determining relation $y = f(C_{5}H_{9}O)$

section S3. Approximation of C$_{5}$H$_{9}O$

section S4. Derivation of Eq. 5 in the main text

section S5. Additional note of Fig. 4 in the main text

section S6. Sensitivity tests regarding entrainment process

section S7. Error analysis

section S8. Comparison of C$_{5}$H$_{9}O$ obtained in this study with other OH studies

section S9. Scatterplots of VOC concentrations with NO$_x$ concentration for all-weather condition

section S10. Simulated daily variation of ozone $\Delta$O$_3$, normalized to the noon time value $\Delta$O$_3$noon, based on Master Chemical Mechanism.

section S11. Simulated NO$_2$ dependence of effective production yields.

section S12. Observation and simulation of ozone concentration C$_O$

section S13. Simulation using a mixed boundary layer model.

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section S15. Variation in model parametrizations for error analysis.

section S16. Production yields and loss rate coefficients for isoprene oxidation products used in the model.

section S17. Uncertainty estimates for inferred C$_{5}$H$_{9}O$noon via error propagation.

section S18. Summary of inferred OH concentrations over tropical forests in South America.

References (41–50)

REFERENCES AND NOTES


28. S. T. Martin, P. Artaxo, L. Machado, A. O. Manzi, R. A. F. Souza, C. Schumacher, J. Wang, T. Biscaro, J. Brito, A. Calheiros, K. Jardine, A. Medeiros, B. Portela, S. S. de Sá, K. Adachi, A. C. Aiken, R. Albrecht, L. Alexander, M. O. Andrehe, H. M. J. Barbosa, P. Buseck, D. Chand, J. M. Comstock, D. Gu, A. B. Guenther, J. E. Shilling, H. Yu, M. Huang, C. Zhao, Q. Yang, S. T. Martin, S. R. Springston, S. T. Martin, P. Artaxo, R. A. F. Souza, K. A. McKinney, S. T. Martin, Isoprene photochemistry over Amazonian Research, and Amazonas State University. We acknowledge the ARM Climate Research Facility, and the São Paulo Research Foundation (2014/05014-0 and 2017/17047-0). Research was conducted under scientific license 001030/2012-4 of the Brazilian National Council for Scientific and Technological Development. Institutional support was provided by the Central Office of the Large-Scale Biosphere-Atmosphere Experiment in Amazonia, the National Institute of Amazonian Research, and Amazonas State University. We acknowledge the ARM Climate Research Facility, a user facility of the DOE, Office of Science, sponsored by the Office of Biological and Environmental Research and support from the Atmospheric System Research program of that office. We thank Y. Li for helpful discussion on error analysis. The research was conducted under scientific license 001030/2012-4 of the Brazilian National Council for Scientific and Technological Development. Funding: Funding was obtained from the DOE (DE-SC0011115 and DE-SC0011122), the NSF (1628491 and 1321987), the Amazonas State Research Foundation, and the São Paulo Research Foundation (2014/05014-0 and 17077). Acknowledgments: Institutional support was provided by the Central Office of the Large-Scale Biosphere-Atmosphere Experiment in Amazonia, the National Institute of Amazonian Research, and Amazonas State University. We acknowledge the ARM Climate Research Facility, a user facility of the DOE, Office of Science, sponsored by the Office of Biological and Environmental Research of the U.S. Department of Energy. The authors declare that they have no competing interests. Data and materials availability: All of the airborne measurement data used in the paper are available at www.arm.gov/campaigns/amf2014goamazon. Additional data related to this paper may be requested from the authors.

Submitted 18 October 2018
Accepted 21 February 2019
Published 11 April 2018
10.1126/sciadv.aar2547

Isoprene photo-oxidation products quantify the effect of pollution on hydroxyl radicals over Amazonia

DOI: 10.1126/sciadv.aar2547