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Sediment Resuspension and Deposition on Seagrass Leaves Impedes Internal Plant Aeration and Promotes Phytotoxic H$_2$S Intrusion

Kasper E. Brodersen¹*, Kathrine J. Hammer², Verena Schrameyer¹, Anja Floytrup², Michael A. Rasheed³, Peter J. Ralph¹, Michael Kühl¹, ⁴ and Ole Pedersen², ⁵*

¹ Climate Change Cluster, Faculty of Science, University of Technology Sydney, Sydney, NSW, Australia, ² Freshwater Biological Laboratory, Department of Biology, University of Copenhagen, Copenhagen, Denmark, ³ Centre for Tropical Water and Aquatic Ecosystem Research (TropWater), James Cook University, Cairns, QLD, Australia, ⁴ Marine Biological Section, Department of Biology, University of Copenhagen, Helsingør, Denmark, ⁵ School of Plant Biology, The University of Western Australia, Crawley, WA, Australia

HIGHLIGHTS:
- Sedimentation of fine sediment particles onto seagrass leaves severely hampers the plants’ performance in both light and darkness, due to inadequate internal plant aeration and intrusion of phytotoxic H$_2$S.

Anthropogenic activities leading to sediment re-suspension can have adverse effects on adjacent seagrass meadows, owing to reduced light availability and the settling of suspended particles onto seagrass leaves potentially impeding gas exchange with the surrounding water. We used microsensors to determine O$_2$ fluxes and diffusive boundary layer (DBL) thickness on leaves of the seagrass Zostera muelleri with and without fine sediment particles, and combined these laboratory measurements with in situ microsensor measurements of tissue O$_2$ and H$_2$S concentrations. Net photosynthesis rates in leaves with fine sediment particles were down to ∼20% of controls without particles, and the compensation photon irradiance increased from a span of 20–53 to 109–145 µmol photons m$^{-2}$ s$^{-1}$. An ∼2.5-fold thicker DBL around leaves with fine sediment particles impeded O$_2$ influx into the leaves during darkness. In situ leaf meristematic O$_2$ concentrations of plants exposed to fine sediment particles were lower than in control plants and exhibited long time periods of complete meristematic anoxia during night-time. Insufficient internal aeration resulted in H$_2$S intrusion into the leaf meristematic tissues when exposed to sediment resuspension even at relatively high night-time water-column O$_2$ concentrations. Fine sediment particles that settle on seagrass leaves thus negatively affect internal tissue aeration and thereby the plants’ resilience against H$_2$S intrusion.

Keywords: diffusive boundary layer, dredging, H$_2$S, in situ, microsensors, photosynthesis, seagrass, sediment
INTRODUCTION

Anthropogenic activities in coastal waters such as dredging, point-source outfall discharges and runoff from agricultural and urban catchments lead to addition and resuspension of fine particulate material that can have substantial negative impacts on the health and fitness of seagrasses (Erfemeijer and Lewis, 2006 and references herein; York et al., 2015; Chartrand et al., 2016). Dredging operations e.g., during harbor expansion or construction work can result in direct removal of plant material and plant burial by suspended sediment. The indirect effects associated with turbid sediment plumes, have largely been attributed to reduced light availability impeding seagrass photosynthesis (e.g., Erfemeijer and Lewis, 2006; York et al., 2015). Dredging-induced seagrass mortality depends on the nature of the dredging operations including the duration and intensity (Erfemeijer and Lewis, 2006; York et al., 2015), but some larger dredging activities result in widespread sediment plumes that can significantly reduce light transmission through the water-column (Cutroneo et al., 2013).

Sediment resuspension for prolonged time periods can strongly affect plant fitness. Even small reductions in light availability can cause pronounced declines in the distribution and growth of seagrass meadows (Ralph et al., 2007; Chartrand et al., 2016). Seagrasses generally have high light requirements and are therefore typically found in waters, where at least 10% of incident solar irradiance reaches the seagrass leaf canopy (Duarte, 1991).

Apart from light attenuation, sediment plumes can also result in the settling of fine sediment particles on seagrass leaves, especially if the plants are already covered by epiphytes that have high potential to trap the sediment e.g., due to their excretion of exopolymers (Pereira et al., 2009; Hamisi et al., 2013). The effects of such sediment coverage on the performance of seagrasses remain largely unexplored although such sediment layers may result in a further substantial reduction in light availability for the underlying leaves, analogous to the adverse shading effects of leaf epiphytes (Brodersen et al., 2015a).

Epiphytic microalgae on seagrass leaves have also been shown to significantly increase the thickness of the diffusive boundary layer (DBL) (Koch, 1994; Brodersen et al., 2015a), that is a thin unstirred layer of water, wherein solute and gas exchange between tissues and the surrounding water occurs by molecular diffusion, which is a slow process compared to bulk exchange of solutes/gasses (e.g., Jorgensen and Revsbech, 1985; Hurd, 2000). The transport time of O$_2$ across the DBL increases with the square of the DBL thickness, i.e., the diffusion path length, and increasing DBL thickness will thus affect the O$_2$ exchange of the seagrass leaf substantially (Jorgensen and Des Marias, 1990; Hurd, 2000; Larkum et al., 2003; Binzer et al., 2005). During the day, thick DBLs may result in increased photorespiration due to tissue accumulation of O$_2$ (e.g., Maberly, 2014), but thick DBLs can be particularly problematic during darkness, where seagrasses completely rely on the diffusive supply of O$_2$ from the surrounding water-column to maintain aerobic respiration of their leaves and below-ground tissues (Borum et al., 2006; Pedersen et al., 2016).

Sediment resuspension may also result in decreased water-column O$_2$ concentrations due to (i) chemical oxidation of reduced metabolites and metals (Erfemeijer and Lewis, 2006), or (ii) increased aerobic mineralization of labile organic matter accumulated in the sediment under anoxic conditions. The chemical and biological O$_2$ demand of suspended particles adds to the substantial O$_2$ consumption by dense seagrass meadows during night-time, potentially resulting in water-column hypoxia (Greve et al., 2003; Borum et al., 2005, 2006). Night-time water-column hypoxia can result in inadequate internal aeration of belowground seagrass tissues resulting in shrinking or disappearance of the oxic micro-shield generated by radial O$_2$ loss (ROL) in the rhizosphere (Koren et al., 2015; Brodersen et al., 2015b). Decreased or absent ROL can result in intrusion of gaseous phytotoxic H$_2$S from the surrounding anoxic sediment into the plant. Once in the plant, the strong binding capacity of H$_2$S with iron in cytochrome c oxidase in the mitochondrial respiratory electron transport chain may inhibit the seagrass metabolism and lead to increased mortality (Raven and Scrimgeour, 1997; Holmer and Bondgaard, 2001; Pérez-Pérez et al., 2012; Lamers et al., 2013). Such H$_2$S intrusion into seagrasses has been demonstrated both under controlled conditions in the laboratory (Pederssen et al., 2004) and in a die-off patch in the field (Borum et al., 2005). Interestingly, seagrasses possess internal detoxification mechanisms, whereby some tissue H$_2$S is oxidized to elemental sulfur within the aerenchyma (Holmer and Hasler-Sheetal, 2014; Hasler-Sheetal and Holmer, 2015). Adequate internal plant aeration is thus a perquisite for healthy seagrass meadows.

The O$_2$ partial pressure (pO$_2$) of seagrass tissues is determined by four main factors: (i) the diffusive O$_2$ flux from the water-column into the leaves during darkness (Pedersen et al., 2004), (ii) photosynthetic O$_2$ production during the day (Dennis, 1987; Fourquerean and Ziemann, 1991), (iii) the respiratory demand of the plant that is strongly affected by the ambient temperature (Raun and Borum, 2013), and (iv) the combined sediment O$_2$ demand affecting the ROL in the rhizosphere (Pedersen et al., 1998; Jensen et al., 2005; Borum et al., 2006; Frederiksen and Gud, 2006).

In the present study, we combined experimental sediment resuspension experiments with microsensor measurements to investigate (i) the rates of photosynthesis and respiration, (ii) the potential role of settled sediment particles on DBL-impedance of O$_2$ exchange with the water-column, (iii) the internal O$_2$ status of the meristematic tissue, and (iv) the meristematic H$_2$S concentrations in the seagrass Zostera muelleri spp. capricorni. Detailed microsensor measurements were performed both under controlled laboratory conditions and in situ, and were coupled to the light, temperature and O$_2$ conditions in the surrounding water-column. We thus tested the hypotheses that sediment deposits on seagrass leaves lead to (i) reduced photosynthetic efficiency, owing to reduced light availability, as well as reduced gas exchange with the surrounding water column, (ii) reduced internal aeration of below-ground seagrass tissue, and (iii) intrusion of H$_2$S into the seagrass. Our data add important ecophysiological information on the
resilience/sensitivity of seagrasses to environmental disturbances linked to anthropogenic activities associated with increases in suspended sediments.

**MATERIALS AND METHODS**

**Seagrass and Sediment Collection**

Specimens of *Z. muelleri* spp. *capricorni* (Asch.) S.W.L. Jacobs and marine sediment were collected from shallow waters (<2 m depth) in Narrabeen Lagoon, NSW, Australia in April 22, 2015. Narrabeen Lagoon is a large (~2 km²), shallow intermittently closed lagoon, with a catchment area of ~55 km². A plastic corer with an inner diameter of 6.3 cm was used to sample bulk sediment cores adjacent to the investigated seagrass meadow. After sampling, seagrasses and sediment were transported to the laboratory, where they were kept in constantly aerated seawater reservoirs (23°C; salinity = 29; mimicking physicochemical water-column conditions at the sampling site) prior to further investigations.

**Sediment Sieving**

Multiple sieves were used to obtain the fine sediment particle fraction with <63 µm grain size, henceforth referred to as silt/clay, from a sheltered area of the lagoon. After sieving, the obtained silt/clay particles and water were left undisturbed overnight in enclosed 10 L containers to allow the suspended particles to resettle. On the following day, the supernatant was carefully drained off avoiding resuspension, and the silt/clay fraction was stored in 1 L sample jars for up to 7 days until used in subsequent experiments. Furthermore, to enable differentiation between physical effects caused by the grains themselves and effects mainly driven by microbial activity within the silt/clay, some of the obtained silt/clay was sterilized by heating it to 120°C in an oven for 2 h within sealed containers to minimize evaporation.

**Laboratory Measurements**

**Experimental Setup**

Leaf segments from 3 randomly selected *Z. muelleri* plants were positioned horizontally in a custom-made flow chamber (Brodersen et al., 2014). Within the chamber, leaf segments were fixed onto a polystyrene plate by needles. The cut ends of the investigated leaf segments were sealed with petroleum jelly prior to experiments to seal the aerenchyma from the surrounding water. A constant flow (~1 cm s⁻¹) of aerated seawater (23°C, salinity = 29) was maintained in the flow chamber via a pump submersed into a seawater reservoir. The applied flow velocity of ~1 cm s⁻¹ is in the lower end of field water velocities (e.g., Gambi et al., 1990; González-Oritz et al., 2014), but does resemble water movement especially within dense seagrass meadows in closed lagoons such as the conditions in Narrabeen Lagoon. Illumination was provided by a fiber-optic tungsten halogen lamp fitted with a collimating lens (KL-2500LCD, Schott GmbH, Germany). The downwelling photon irradiance (PAR, 400–700 nm) at the leaf surface was measured with a scalar irradiance minisensor (US-SQS/L, Walz GmbH, Germany) connected to a calibrated photon irradiance meter (LI-250A, LI-COR, USA). The leaf segments were illuminated with an incident photon irradiance of 0, 75, 200, and 500 µmol photons m⁻² s⁻¹. Water-column hypoxia was obtained by continuously flushing the seawater in the supporting water reservoir with a mixture of atmospheric air and humidified nitrogen gas. The O₂ concentration of the water reservoir was simultaneously monitored by a submerged Clark-type O₂ microsensor (OX-10, tip diameter of 10 µm, Unisense A/S, Aarhus, Denmark; Revsbech, 1989).

**O₂ Microsensor Measurements**

We used Clark-type O₂ microsensors (OX-50, tip diameter of ~50 µm, detection limit ~0.3 µM, Unisense A/S, Aarhus, Denmark; (Revsbech, 1989)) with a fast response time (t₉₀ <0.5 s) and a low stirring sensitivity (<2–3%) to measure the O₂ concentration at and toward the leaf surface. The O₂ microsensors were mounted on a motorized micromanipulator (Unisense A/S, Aarhus, Denmark) and connected to a microsensor multimeter (Unisense A/S, Aarhus, Denmark) both interfaced with a PC running dedicated data acquisition and positioning software (SensorTrace Pro, Unisense A/S, Aarhus, Denmark). The O₂ microsensors were linearly calibrated from signal readings in 100% air saturated seawater and anoxic seawater (by N₂ flushing and addition of the O₂ scavenger Na₂SO₃) at experimental temperature and salinity. Prior to measurements and calibrations, the microsensors were preconditioned with H₂S to prevent drifting calibrations when exposed to H₂S during experiments (Brodersen et al., 2015a). Microsensors were carefully positioned at the leaf tissue surface (defined as 0 µm distance on figures) by manually operating the micromanipulator, while observing the leaf tissue surface and microsensor tip with a boom-stand dissection microscope (AmScope, Irvine, CA, USA). When changing the downwelling photon irradiance, steady state O₂ conditions at the leaf surface re-occurred after ~60 min (data not shown). Microprofiles of O₂ concentration were measured in vertical increments of 100 µm, from the leaf tissue surface to 2 mm distance away (which is in the same order of magnitude as the leaf tissue thickness).

**Photosynthesis and Respiration Calculations**

O₂ fluxes across the leaf tissue surfaces were calculated using Fick’s first law of diffusion:

\[ \mathbf{J}_{O_2} = -D_{O_2} \frac{\partial C}{\partial z} \]  

(1)

where \( D_{O_2} \) is the molecular diffusion coefficient of O₂ in seawater at experimental temperature and salinity (2.14 × 10⁻⁵ cm² s⁻¹; cf. tabulated physical parameters for marine systems available at www.unisense.com), and \( \partial C/\partial z \) is the linear O₂ concentration gradient in the DBL. As we introduced a physical barrier to O₂ diffusion at the abaxial surface by fixing the leaf onto polystyrene with a low O₂ permeability, we take the flux estimated at the adaxial side of the seagrass leaf as representative for the net flux of O₂ across the leaf surface, i.e., \( J_{O_2, tot} = J_{O_2, upper-surface} \) in dark (=respiration) and light (=net photosynthesis; assuming a photosynthetic quotient of 1 mol O₂ produced per mol CO₂ fixed), respectively.
The calculated net photosynthesis rates (nmol O$_2$ m$^{-2}$ s$^{-1}$) as a function of the incident photon irradiance (E; µmol photons m$^{-2}$ s$^{-1}$) were fitted with an exponential saturation model (Webb et al., 1974) with an added term, R, to account for respiration (Spilling et al., 2010):

$$P_0(E) = P_{max} \left(1 - e^{-\frac{E}{P_{max}}}\right) + R \quad (2)$$

This equation enables estimation of the irradiance at the onset of photosynthesis saturation as $E_k = P_{max}/\alpha$, where $P_{max}$ is the maximal net photosynthesis rate and $\alpha$ is the initial slope of the $P_0$ vs. $E$ curve. The compensation photon irradiance, $E_C$, was determined as the incident photon irradiance at which the leaf tissue shifted from a net O$_2$ consumption to a net O$_2$ production, i.e., the photon irradiance where $P_0(E) = 0$.

### Bulk Sediment O$_2$ Uptake

Depth profiles of O$_2$ concentration in the bulk sediment were obtained as follows. The sediment core was submerged into a ~2 L aquarium, wherein stirring and aeration of the water column was achieved via a Pasteur pipette connected to an air-pump. The surface of the sediment was determined with a boom-stand dissection microscope (AmScope, Irvine, CA, USA) and the O$_2$ microsensors were carefully positioned at the sediment surface as described above. Microprofiles were performed in vertical increments of 200 µm down to 2 cm depth, i.e., below the O$_2$ penetration depth. The volume specific O$_2$ consumption rate of the bulk sediment, $R_{sed}$ (µmol O$_2$ m$^{-3}$ s$^{-1}$), was calculated as:

$$R_{sed} = \frac{I_{O_2}}{d_{O_2}} \quad (3)$$

where $I_{O_2}$ is the O$_2$ flux at the seawater/sediment interface (µmol O$_2$ m$^{-2}$ s$^{-1}$), i.e., the diffusive oxygen uptake (DOU) of the sediment as calculated from Equation (1), and $d_{O_2}$ is the O$_2$ penetration depth in the sediment (cm) as shown in Figure S1 (Supplementary Materials).

### Potential and Biological O$_2$ Consumption of Sieved Sediment

The O$_2$ consumption of the fine sediment particles used in the laboratory as well as in situ was determined using a slightly modified approach of Pedersen et al. (2011). The O$_2$ consumption was separated into total (OX$_{tot}$) or biological (OX$_{bio}$) O$_2$ demand in order to determine the chemical O$_2$ demand as OX$_{chem} = OX_{tot} - OX_{bio}$. The total O$_2$ consumption of the sediment fraction was determined by mixing 50 mL suspended sediment (~63 µm) with 950 mL seawater with a salinity of 28. The solution was immediately transferred into 25 mL glass vials fitted with 2 glass beads to provide mixing and mounted on a rotating wheel (8 rpm) in a constant temperature bath (20.0 ± 0.5°C) (Pedersen et al., 2013). The sediment suspension was incubated for about 1 h (exact times recorded) before the O$_2$ concentration was measured in each vial using a calibrated sturdy O$_2$ microsensor (OX500; Unisense A/S, Denmark). Vials with seawater but without suspended sediment served as blanks enabling calculation of the O$_2$ consumption as µmol O$_2$ m$^{-3}$ sediment s$^{-1}$.

The biological O$_2$ consumption was measured on a sediment suspension, which was initially purged with atmospheric air for 15 min to oxidize reduced metals and sulfide (Raun et al., 2010). After oxidation, the sediment suspension was transferred into 25 mL glass vials and treated as described above.

### In situ Measurements

#### Experimental Setup

Two patches (~1 m in diameter) of Z. muelleri were enclosed by custom-made transparent, floating curtains with mixing provided by submerged pumps to simulate water motion outside the enclosures (Narrabeen Lagoon, Australia). One enclosure functioned as a control treatment and the other enclosure as a silt/clay treatment. In the silt/clay treatment, 3 pulses of 375 mL silt/clay particles (see above) were added to the water column per day to mimic a dredging operation. Sediment resuspension was initiated at the beginning of the experiments (afternoon) (pulse 1), just before sunrise (pulse 2) and at midday (pulse 3). Measurements were performed on April 17, 2015 (Series 1) and repeated on April 19, 2015 (Series 2), i.e., there were 27 h difference between Series 1 and Series 2 measurements. Within the enclosures, we measured salinity, light, temperature and O$_2$ in the water column during measurements of meristematic tissue O$_2$ and H$_2$S concentrations. A detailed description of the in situ measurements is given below.

#### Internal $pO_2$ and [H$_2$S] Measurements

Similar data acquisition equipment and microsensor as described above were used for the field measurements of internal O$_2$ partial pressure ($pO_2$) and H$_2$S concentrations ([H$_2$S]) in the meristematic tissue of Z. muelleri over diel cycles. Internal H$_2$S concentrations were measured with Clark-type H$_2$S microsensors (H2S-25, tip diameter of ~25 µm, 90% response time <10 s, detection limit ~0.3 µm, Unisense A/S, Aarhus, Denmark; Jeroschewski et al., 1996; Küh, et al., 1998) that were linearly calibrated in anoxic, acidic (pH 4) Na$_2$S solutions of known H$_2$S concentrations (0, 50, and 100 µM).

Within the enclosures, the microsensors were mounted on micromanipulators that were supported by stabilized aluminum spears at a water depth of ~1 m. The O$_2$ and H$_2$S microsensors were simultaneously inserted into the briefly-exposed shoot base of the target plants close to the basal leaf meristem, which was then re-buried ~2 cm into the sediment to re-establish the biogeochemical gradients (Pedersen et al., 2004). Positioning of the O$_2$ microsensors was done by observing the sensor signals during insertion until a constant signal was recorded (Borum et al., 2005). The H$_2$S microsensors were inserted via a similar approach, using a combination of sensor signal responses to light exposure and positioning the electrodes at approximately the same depth into the leaf meristem tissue as the O$_2$ microsensors. The intra-plant O$_2$ and H$_2$S concentrations were measured simultaneously inside one plant in the control treatment and one plant in the silt/clay treatment, and then replicated.
Physical and Chemical Parameters of the Water-Column

Diel changes in ambient incident photon irradiance (continuously measured via Odyssey light loggers; Dataflow Systems, Christchurch, NZ), water-column $pO_2$ (via $O_2$ micro-optodes; OXF500PT, Pyroscience, Aachen, Germany; connected to a 4-channel Firesting meter, PyroScience, Germany), and water-column temperature (via HOBO temperature data loggers; UA-002-08, Onset Computer Corporation, Bourne, MA, USA) were recorded over $\sim$24 h within the enclosures. All sensors were calibrated according to the manufactures instructions, mounted on a metal spear and positioned at leaf canopy height. Logging (1 Hz) by all data loggers was synchronized with the logging of microsensors used for the intra-tissue measurements.

In situ calculations

All microsensors are temperature sensitive (e.g., Kühl and Revsbech, 2001) and thus the measurements of internal $pO_2$ and [H$_2$S] obtained by the calibrated $O_2$ and H$_2$S microsensors were temperature corrected using the following equations (available at www.unisense.com):

$$pO_2 = \frac{S_{amb} - Z}{S_{air} - Z} P_0 e^{k(T_{cal} - T_{amb})}$$  \hspace{1cm} (4)

where $S_{amb}$ is the sensor signal measured in situ (mV), $S_{air}$ is the calibration signal of the sensor determined at known partial pressure and temperature (e.g., 100% air saturation; in mV), $Z$ is the zero current of the sensor measured at known partial pressure and temperature (i.e., 0% air saturation; in mV), $P_0$ is the known partial pressure used to define $S_{air}$ (kPa), $k$ is the temperature coefficient of the respective sensor ($\sim$0.02°C$^{-1}$; exact values for individual sensors can be provided by the manufacturer, www.unisense.com), $T_{cal}$ is the known calibration temperature ($^\circ$C), and $T_{amb}$ is the ambient temperature ($^\circ$C) continuously measured in situ.

$$[H_2S] = (G + S_0) e^{k(T_{cal} - T_{amb})}$$  \hspace{1cm} (5)

where $G$ is the slope of the calibration curve that represents the sensitivity of the sensor ($\mu$mol L$^{-1}$ mV$^{-1}$), $S$ is the signal of the sensor (mV), $S_0$ is a constant that describes the zero current ($\mu$mol L$^{-1}$), $k$ is the temperature coefficient of the respective sensor ($\sim$0.02°C$^{-1}$), $T_{cal}$ is the known calibration temperature ($^\circ$C), and $T_{amb}$ is the ambient temperature ($^\circ$C) continuously determined in situ.

These final sensor calibrations were done after the in situ experiments using the temperature data obtained in the respective enclosures by the submerged HOBO temperature data loggers (HOBO, Onset Computer Corporation, Bourne, MA, USA).

Data Analysis

In the following, $O_2$ is quantified as $\mu$mol L$^{-1}$ when in solution and as kPa when in gaseous phase. Data obtained under controlled conditions in the laboratory, i.e., $O_2$ fluxes across the leaf tissue surface, are thus presented in molar concentrations and data obtained in situ, i.e., meristematic $O_2$ concentrations and water-column $O_2$ conditions are given as partial pressures. Furthermore, all laboratory measurements were performed at 40 and 100% air equilibrium, representing water-column $O_2$ conditions at night- and day-time, respectively. Non-linear curve fitting was used to estimate the relationship among variables. All data fitting and analyses were performed in OriginPro (OriginPro 8, OriginLab Corporation, Northampton, MA, USA).

RESULTS

Laboratory Measurements

Sediment and Silt $O_2$ Consumption Rates

To enable comparison of sediment activity, we determined the $O_2$ demand and characteristics of the added silt/clay particles ($<63$ μm) and the bulk sediment without seagrass biomass. The $O_2$ was depleted within the upper 1.2 mm of the bulk sediment and the sediment remained anoxic with depth (Figure S1). The volume-specific $O_2$ consumption rate of the bulk sediment was estimated to $374 \pm 33 \mu$mol $O_2$ m$^{-3}$ s$^{-1}$ (Table 1). In contrast, the fine sediment particles consumed $1319 \pm 6 \mu$mol $O_2$ m$^{-3}$ s$^{-1}$ when taking both the biological and chemical $O_2$ demand into account. The biological $O_2$ demand of the silt/clay particles was $1254 \pm 29 \mu$mol $O_2$ m$^{-3}$ s$^{-1}$ resulting in a chemical $O_2$ demand of $65 \mu$mol $O_2$ m$^{-3}$ s$^{-1}$ (Table 1). Hence, the chemical $O_2$ demand of the fine sediment particles can thus most likely be neglected.

Net Photosynthesis and Respiration Rates

Net photosynthesis rates increased with increasing incident photon irradiance for both plants with and without leaf silt/clay-cover (Figure 1; showing $O_2$ fluxes from/into leaves). Moreover, net photosynthesis rates were higher in control leaf segments (no silt/clay added) exposed to hypoxic water conditions, resembling water-column $O_2$ levels at sunrise, as compared to leaf segments kept in water at 100% air equilibrium (Table 2). Plants with leaf silt/clay-cover exhibited net $O_2$ consumption already at an incident photon irradiance of $\sim$75 $\mu$mol photons m$^{-2}$ s$^{-1}$ owing to reduced light availability for leaf photosynthesis (Figure 1; Table 2). Net photosynthesis rates of the control plants were 3 to 5-fold higher under moderate photon irradiance (200 $\mu$mol photons m$^{-2}$ s$^{-1}$) as compared to plants with leaf silt/clay-cover.

<table>
<thead>
<tr>
<th>Sediment type</th>
<th>$O_2$ consumption ($\mu$mol m$^{-3}$ s$^{-1}$)</th>
</tr>
</thead>
<tbody>
<tr>
<td>Bulk, sediment ($P_{sed}$)</td>
<td>374 ± 33</td>
</tr>
<tr>
<td>Fine sediment particles (Biological $O_2$ demand, $OX_{bio}$)</td>
<td>1254 ± 29</td>
</tr>
<tr>
<td>Fine sediment particles (Biological and chemical $O_2$ demand, $OX_{total}$)</td>
<td>1319 ± 6</td>
</tr>
</tbody>
</table>

Rates are mean values ± SE; n = 4. Biological $O_2$ demand refers to the $O_2$ consumption of fine sediment particles oxygenated via 15 min air flushing prior to measurements. Biological and chemical $O_2$ demand of fine sediment particles refers to the $O_2$ consumption rate of untreated, i.e., not purged with air prior to incubation, fine sediment particles. 50 mL fine sediment particles were added to 950 mL seawater.
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FIGURE 1 | Vertical O$_2$ concentration profiles measured toward the leaf surface under incident photon irradiances of 0, 75, 200, and 500 µmol photons m$^{-2}$ s$^{-1}$. Red symbols and lines represent leaves with silt/clay-cover; black symbols and lines represent control plants, i.e., leaves without silt/clay-cover. Upper panels are measurements in water with a reduced O$_2$ level of ~40% of air equilibrium (mimicking night-time water-column O$_2$ conditions, approximately 8.2 kPa); lower panels are measurements in water at 100% air equilibrium (mimicking day-time water-column O$_2$ conditions, 20.6 kPa). Zero depth indicates the leaf surface. Symbols and error bars represent means ± SE; n = 3–4.

TABLE 2 | Gas exchange measured as the O$_2$ flux across leaf surfaces of plants without (control)- and with fine sediment particles (<63 µm) as a function of photon irradiance.

<table>
<thead>
<tr>
<th>Downwelling photon irradiance (µmol photons m$^{-2}$ s$^{-1}$)</th>
<th>Control 40% air equilibrium (nmol O$_2$ m$^{-2}$ s$^{-1}$)</th>
<th>With fine sediment particles 40% air equilibrium (nmol O$_2$ m$^{-2}$ s$^{-1}$)</th>
<th>Control 100% air equilibrium (nmol O$_2$ m$^{-2}$ s$^{-1}$)</th>
<th>With fine sediment particles 100% air equilibrium (nmol O$_2$ m$^{-2}$ s$^{-1}$)</th>
</tr>
</thead>
<tbody>
<tr>
<td>0</td>
<td>−205 ± 57</td>
<td>−132 ± 3</td>
<td>−663 ± 223</td>
<td>−479 ± 44</td>
</tr>
<tr>
<td>75</td>
<td>435 ± 148</td>
<td>−18 ± 47</td>
<td>179 ± 61</td>
<td>−84 ± 143</td>
</tr>
<tr>
<td>200</td>
<td>854 ± 342</td>
<td>164 ± 110</td>
<td>571 ± 274</td>
<td>195 ± 129</td>
</tr>
<tr>
<td>500</td>
<td>746 ± 143</td>
<td>270 ± 74</td>
<td>701 ± 217</td>
<td>481 ± 266</td>
</tr>
</tbody>
</table>

Positive values denote O$_2$ efflux across the seagrass leaf surface. Rates are mean ± SE; n = 3–4. Note that the relative high standard errors in the silt treatment at 75 µmol photons m$^{-2}$ s$^{-1}$ was due to one of the leaf segments producing O$_2$ via photosynthesis (for further information, please see Figure S2).

(Table 2). During darkness, a constant diffusive O$_2$ influx across the leaf surfaces of both plants with and without leaf silt/clay-cover was observed (Figure 1). However, we found a reduction in the O$_2$ flux into the silt/clay-covered leaves of 28–35% as compared to leaves without silt/clay-cover (Table 2; measured at 100% and 40% air equilibrium, respectively).

During water-column hypoxia, the leaf silt/clay-layer impeded the diffusive O$_2$ supply resulting in almost anoxic conditions at the leaf tissue surface (~16 µmol O$_2$ L$^{-1}$) of plants with leaf silt/clay-cover. This substantially increased the risk of H$_2$S intrusion into the below-ground tissues during night-time as a result of inadequate internal aeration (Figure 1). The thickness of the DBL surrounding the leaves increased from ~200 µm to ~500 µm in the presence of the leaf silt/clay layer (Figure 2). This resulted in a reduction in the O$_2$ influx to the leaves from 484 ± 133 nmol O$_2$ m$^{-2}$ s$^{-1}$ in plants without leaf silt/clay-cover to 419 ± 145 nmol O$_2$ m$^{-2}$ s$^{-1}$ in plants with an inactivated leaf silt/clay-layer. When coated with a biologically active silt/clay layer, leaves exhibited a further reduction of the O$_2$ influx to 395 ± 102 nmol O$_2$ m$^{-2}$ s$^{-1}$ (Figure 2).

The silt/clay-cover on seagrass leaves resulted in a pronounced increase of the plants’ compensation irradiance from 53 ± 7 µmol photons m$^{-2}$ s$^{-1}$ for control leaf segments to 145 ± 46 µmol photons m$^{-2}$ s$^{-1}$ for leaf segments with silt/clay cover, both kept in a water column at 100% air equilibrium (Figure 3; Table 3). In a water column with O$_2$ kept at 40% atmospheric
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**FIGURE 2** | Vertical depth profiles of the O$_2$ concentration measured toward the leaf surface of plants with a microbially active silt/clay-cover (red symbols and lines), with an inactivated silt/clay-cover (obtained by pre-heating the added silt/clay to 120°C in an oven for 2 h; blue symbols and lines), and without silt/clay-cover (control plants; black symbols and lines). All measurements were performed in darkness. Zero depth indicates the leaf surface. The effective DBL thickness can be estimated by extrapolating the linear O$_2$ concentration gradient until it intersects with the constant O$_2$ concentration in the overlying water. The distance from this point into the leaf tissue surface is a measure of the effective DBL thickness (Jørgensen and Revsbech, 1985). Symbols and error bars represent means ± SE; n = 4.

In situ Measurements and Effects of Sediment Re-suspension

**FIGURE 3** | Apparent net photosynthesis rates as a function of downwelling photon irradiance (PAR, 400–700 nm) of plants with leaf silt/clay-cover (red symbols and lines) and without leaf silt/clay-cover (control plants; black symbols and lines). Rates were calculated for incident photon irradiances of 0, 75, 200, and 500 μmol photons m$^{-2}$ s$^{-1}$ and were fitted with an exponential function (Webb et al., 1974) with an added term to account for respiration (Spilling et al., 2010) ($R^2_{40\% \text{ AE, control}} = 0.93$; $R^2_{40\% \text{ AE, silt/clay-cover}} = 0.98$; $R^2_{100\% \text{ AE, control}} = 0.99$; $R^2_{100\% \text{ AE, silt/clay-cover}} = 0.99$). The upper panel represents measurements in water kept at 40% air equilibrium, while the lower panel represents measurements in water kept at 100% air equilibrium. Error bars are ± SE; n = 3–4.

Diel Changes in the Physical/Chemical Parameters of the Surrounding Water-Column

The pO$_2$ dynamics in the water-column of the control and silt/clay treatment showed similar patterns on a diel basis, with steadily declining pO$_2$ during night-time reaching minimal water-column O$_2$ conditions around sunrise, followed by a rapid increase in the water-column pO$_2$ shortly after sunrise approaching atmospheric saturation (20.6 kPa) or even leading to water-column supersaturation relative to atmospheric pO$_2$ around midday (Figures 4A,B). Water-column O$_2$ levels within the enclosures fluctuated substantially during night-time owing to water bodies with varying O$_2$ content being introduced to the seagrass meadow from non-vegetated areas within the lagoon.
TABLE 3 | Photosynthetic parameters derived from the light response curves in Figure 3.

<table>
<thead>
<tr>
<th></th>
<th>40% of air equilibrium</th>
<th>In air equilibrium</th>
</tr>
</thead>
<tbody>
<tr>
<td></td>
<td>Control</td>
<td>Fine sediment particles</td>
</tr>
<tr>
<td>( \alpha )</td>
<td>15 ± 4</td>
<td>3 ± 1</td>
</tr>
<tr>
<td>( P_{\text{max}} )</td>
<td>1028 ± 176</td>
<td>503 ± 91</td>
</tr>
<tr>
<td>( R )</td>
<td>−211 ± 48</td>
<td>−141 ± 4</td>
</tr>
<tr>
<td>( E_C )</td>
<td>20 ± 8</td>
<td>109 ± 47</td>
</tr>
<tr>
<td>( E_S )</td>
<td>72 ± 5</td>
<td>174 ± 46</td>
</tr>
<tr>
<td>([O_2], \text{dark} )</td>
<td>59 ± 14</td>
<td>16 ± 14</td>
</tr>
</tbody>
</table>

Including photosynthetic activity, compensation irradiance, onset of photosynthesis saturation and respiration rates of investigated Zostera muelleri spp. capricorni plants with- and without (i.e., control plants) fine sediment particles on leaves. All photosynthetic related parameters were determined at both 40% of air equilibrium and in air equilibrium. \( n = 3 \). Values are mean ± SE. \( \alpha \) = initial slope of the net photosynthesis rate vs. incident photon irradiance; \( P_{\text{max}} \) = maximum rate of net photosynthesis (in \( \text{nmol} \text{O}_2 \text{m}^{-2} \text{s}^{-1} \)); \( R \) = the respiration rate (in \( \text{nmol} \text{O}_2 \text{m}^{-2} \text{s}^{-1} \)); \( E_C \) = compensation irradiance (i.e., incident photon irradiance where the oxygen produced by photosynthesis meets the respiratory demands) (in \( \mu \text{mol photons} \text{m}^{-2} \text{s}^{-1} \)); \( E_S \) = onset of photosynthesis saturation (in \( \mu \text{mol photons} \text{m}^{-2} \text{s}^{-1} \)); \([O_2], \text{dark} \) = the leaf surface \( \text{O}_2 \) concentration measured in darkness (in \( \mu \text{mol} \text{L}^{-1} \)), which can be used as an estimate for the internal \( \text{O}_2 \) concentration in the aerenchymal tissue of the thin seagrass leaves. 40% of air equilibrium mimics natural conditions in the seagrass meadow during night-time and at sunrise as seen on Figure 4. Air equilibrium mimics natural conditions during most of the day-time (Figure 4). Values are calculated/extracted from the fitted exponential saturation function (Webb et al., 1974) with an added term to account for respiration (Spilling et al., 2013) in Figure 3 (apply to: \( a, P_{\text{max}}, R, E_C \), and \( E_S \)) and from the \( \text{O}_2 \) concentration microprofiles in Figure 1 (\([O_2], \text{dark} \)); and thus all originate from the laboratory experiments.

and/or from the ocean due to tidal water movement. In contrast, water-column temperature remained relatively constant on a diel basis but generally decreased from \( \sim 22^\circ \text{C} \) on the first measuring day (Series 1) to \( \sim 20^\circ \text{C} \) at the end of the second measuring day (Series 2). Minor fluctuations in the water-column temperature during night-time were correlated with the passing of aerated water bodies as observed in the water-column \( \text{O}_2 \) measurements (Figures 4A,B). The incident photon irradiance measured at leaf canopy height followed a typical bell-shaped diurnal curve, with minor fluctuations in the control treatment due to passing cloud cover. This was in strong contrast to the silt/clay treatment, where we measured substantially reduced light conditions as compared to the control treatment, especially in the hours following experimentally manipulated silt/clay re-suspension (Figures 4A,B). Moreover, a pronounced difference in the light availability was observed between measuring days Series 1 and Series 2, where Series 1 represented sunny conditions and Series 2 represented a cloudy late autumn day at Narrabeen Lagoon (Figures 4A,B).

**In situ Measurements of \( \text{O}_2 \) and \( \text{H}_2\text{S} \) in Seagrass Meristems**

The internal, meristematic \( \text{pO}_2 \) of both control plants and plants experimentally exposed to suspended silt/clay decreased steadily from early in the afternoon throughout the night. A minimum internal, meristematic \( \text{pO}_2 \) was reached shortly after sunrise. Thereafter, a rapid increase in meristematic \( \text{pO}_2 \) occurred as a response to increasing solar irradiance resulting in photosynthetic \( \text{O}_2 \) production (Figures 4C,D). Control plants as well as silt/clay-treated plants exhibited lower \( \text{pO}_2 \) relative to the water-column during night-time with tissue \( \text{pO}_2 \) fluctuations correlating with changes in water-column \( \text{pO}_2 \) (Figures 4A-D).

A clear discrepancy in the meristematic \( \text{pO}_2 \) between control plants and leaf silt/clay-treated plants was measured during light-limitation in the early morning hours (06:30–09:00) (Figure 4C) with relatively lower \( \text{pO}_2 \) in silt/clay-treated plants indicating a silt/clay-induced reduction in light availability.

The meristematic below-ground tissues of both control and silt/clay-treated plants turned anoxic, or severely hypoxic, late at night. Meristematic \( \text{pO}_2 \) of silt/clay-treated plants reached anoxia from around 05:00–06:30 in Series 1 and already from 23:30 in Series 2, while the control plants only were exposed to anoxic conditions in the meristematic tissue for short time periods (<1 h; Figures 4C,D). Simultaneous measurements of internal, meristematic \( \text{H}_2\text{S} \) concentrations revealed phytotoxic \( \text{H}_2\text{S} \) intrusion into silt/clay-treated plants during night-time in Series 2 from around 23:30 correlating with the recorded period of meristematic tissue anoxia (Figures 4C,D). Internal \( \text{H}_2\text{S} \) levels reached a maximum of 8.3 \( \mu \text{mol} \text{H}_2\text{S} \text{L}^{-1} \) around 08:00 in the morning and then started to decrease shortly after sunrise in response to photosynthetic \( \text{O}_2 \) production leading to disappearance of \( \text{H}_2\text{S} \) in the meristem by 10:30. No \( \text{H}_2\text{S} \) intrusion was detected into the control plants.

**Effects of Water Column \( \text{O}_2 \) Levels and Silt/Clay on Internal \( \text{O}_2 \) Status**

During night-time, tissue \( \text{pO}_2 \) was derived from \( \text{O}_2 \) in the surrounding water diffusing into the leaves and spreading via aerenchyma to below-ground tissues (Pedersen et al., 1998; Colmer, 2003; Brodersen et al., 2015a). The critical water column \( \text{O}_2 \) level was defined as the water column \( \text{pO}_2 \) below which toxic conditions in the meristematic tissue could no longer be sustained, and this critical \( \text{O}_2 \) level was estimated by plotting the internal \( \text{pO}_2 \) determined in situ against water-column \( \text{pO}_2 \) (Figure 5). In Series 1, the meristematic tissue of the silt/clay-treated plant became anoxic at a water-column \( \text{pO}_2 \) of \( \sim 5.5 \text{ kPa} \) during night-time as compared to \( \sim 8.7 \text{ kPa} \) in the control plant (Figures 5A,C); a tendency that dramatically changed during prolonged exposure to suspended silt/clay particles (i.e., in Series 2) where the silt/clay-treated plant became anoxic already at a night-time water-column \( \text{pO}_2 \) of \( \sim 13 \text{ kPa} \) as compared to \( \sim 6.4 \text{ kPa} \) in the control plant (Figures 5B,D). These in situ findings aligned well with the lower \( \text{O}_2 \) influx into leaves with silt/clay-cover, as compared to control leaves, determined in the controlled laboratory experiments during darkness (Figures 1–3; Tables 2,3).

The silt/clay-induced shading effects on the intra-plant \( \text{pO}_2 \) during natural light exposure of the seagrass leaf canopy was evaluated by plotting the in situ meristematic \( \text{pO}_2 \) as a function of incident photon irradiance (Figure 6) revealing an \( \sim 45\% \) reduction in meristematic \( \text{pO}_2 \) in plants exposed to suspended silt/clay as compared to control plants, seen as a decrease in \( \alpha \), i.e., the slope describing the internal \( \text{O}_2 \) evolution as a function of photon irradiance, from 0.14 to 0.08 (Figure 6).

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DISCUSSION

Our results provide strong evidence that silt/clay-cover on seagrass leaves can have substantial negative effects on the plants’ photosynthetic activity and efficiency, as well as on the night-time O₂ exchange between leaf tissue and the surrounding water. Reduced internal aeration, and thus decreased below-ground tissue oxidation capacity, rendered plants with leaf silt/clay-cover more prone to H₂S intrusion even at relatively high water-column pO₂ during night-time. Below, we discuss in detail the implications of reduced light availability for photosynthesis owing to silt/clay shading, thicker DBLs, and the introduction of O₂ consumption within the DBL itself, on internal aeration and whole plant performance of seagrasses.

Sediment and Silt/Clay Characteristics

We measured an ~3.4-fold higher volumetric O₂ consumption rate of the fine sediment particles (~63 µm), as compared to the bulk sediment, indicative of high microbial activity within the thin silt/clay layer covering the leaf (Table 1). Microbial O₂ respiration was the quantitatively most important O₂ consuming process of the fine sediment particles, while chemical oxidation only accounted for ~5% of the total O₂ demand (Table 1). Hence, the leaf silt/clay-cover not only impeded gas and nutrient exchange with the surrounding water-column owing to the enhanced thickness of the DBL around the leaves (Figure 2), it also reduced the passive O₂ influx across the silt/clay layer during night-time owing to high microbial O₂ consumption within the silt/clay layer.
In light, the apparent net photosynthesis rates of *Z. muelleri* Photosynthesis and O$_2$ uptake were greatly reduced as compared to control leaves, and the reduction was most pronounced at low to moderate photon irradiances (Figure 3, Table 2). The reduced leaf photosynthesis was, most likely, a combined negative result of lower light availability at the tissue surface and a DBL-impeded uptake of CO$_2$ from the surrounding water-column, potentially leading to enhanced photorespiration and thereby reduced photosynthetic efficiency owing to inorganic carbon limitation (e.g., Maberly, 2014; Figures 1, 3, Table 2). Consequently, the compensation irradiance of photosynthesis for plants with leaf silt/clay-cover increased to $\sim$109 and 145 $\mu$mol photons m$^{-2}$ s$^{-1}$, as compared to $\sim$20 and 53 $\mu$mol photons m$^{-2}$ s$^{-1}$ for control plants kept in water with O$_2$ at 40% air equilibrium and 100% air equilibrium, respectively (Table 3). Silt/clay-cover can thus keep seagrass plants close to their minimum light requirements on days with poor light conditions. However, in the present experimental set-up we were unable to clearly separate the effect of reduced net photosynthesis caused by reduced light (shading by particles) or increased resistance to CO$_2$ influx (thicker DBL) from that of O$_2$ consumption by bacteria within the silt/clay layer.

In darkness, the passive O$_2$ influx was also strongly affected by the leaf silt/clay-cover, causing a reduction of up to 35% in the O$_2$ supply (Table 2), which resulted in reduced internal aeration (Table 3) especially under hypoxic water-column conditions and thus markedly increased the risk of over-night tissue anoxia. The lower O$_2$ influx was a combined negative result of an increased DBL thickness impeding the exchange of O$_2$ with the surrounding water-column (Figure 2) and high microbial O$_2$ consumption rates within the leaf silt/clay-cover (Figure 2, Table 1). Such reduction in the meristematic pO$_2$ lead to a reduced capability of the silt/clay-covered seagrass plant to aerate its below-ground tissue during night-time increasing the risk for phytotoxic H$_2$S intrusion (Pedersen et al., 2004; Borum et al., 2005; Brodersen et al., 2015b).

Moreover, at high irradiances the silt/clay-induced impeded gas exchange with the water column may also lead to supra-optimal internal O$_2$ levels during daytime, potentially resulting in oxidative stress or damage (Brodersen et al., 2015a). Similarly, elevated temperatures may lead to a CO$_2$ build-up at night-time that could result in a negative feedback on respiration, cellular pH and rates of dark fixation.

**Sediment Re-suspension Effects on Plant Meristematic O$_2$ and H$_2$S Levels**

Diel changes in the meristematic O$_2$ content of seagrasses were mainly driven by irradiance (Figure 4). Experimentally manipulated silt/clay re-suspension within the enclosure of the silt/clay treatment, resulted in a pronounced decrease of light availability for seagrass photosynthesis with up to 3 h of darkening measured around midday in Series 2 (Figure 4B). The diminished light conditions resulted in reduced photosynthetic O$_2$ evolution and thereby reduced meristematic pO$_2$ in *Z. muelleri* as seen at sunrise in Series 1 (Figure 4C), thus correlating with previous findings by Borum et al. (2005). The photosynthetic efficiency of *Z. muelleri* measured in situ was also strongly affected by the silt/clay exposure, with an almost
2-fold decrease in the net photosynthetic O\textsubscript{2} evolution of plants exposed to fine sediment particles, as compared to control plants at equivalent incident photon irradiances (Figure 6), leading to reduced internal aeration and below-ground tissue oxidation capacity. This was a result of impeded gas exchange with the surrounding water-column due to a thicker DBL in the presence of a sediment cover of leaves leading to lower photosynthetic efficiencies. The in situ measurements thus strongly correlated with findings of a 3–5-fold higher compensation irradiance and an \( \sim 2.4 \)-fold increase in the irradiance at onset of photosynthesis saturation in the laboratory experiments for Z. muelleri leaves with silt/clay-cover as compared to control leaves (Figure 3; Table 3).

Critically low meristematic pO\textsubscript{2} and/or tissue anoxia were only measured during night-time and occurred for longer periods of time, and at higher water-column O\textsubscript{2} levels, for Z. muelleri in the silt/clay treatment as compared to the control treatment (Figures 4, 5). This suggests reduced O\textsubscript{2} supply from the leaves to the below-ground tissue of Z. muelleri plants exposed to fine sediment particles. The reduced meristematic pO\textsubscript{2} was caused by (i) the leaf silt/clay-cover induced enhanced DBL thickness impeding the passive O\textsubscript{2} influx into the leaves, and (ii) high rates of microbial O\textsubscript{2} consumption within the leaf silt/clay layer in line with observations in the laboratory experiments (Figures 1,2). Lowest meristematic pO\textsubscript{2} levels were recorded around sunrise, followed by a rapid increase in the meristematic O\textsubscript{2} content when sunlight supported leaf photosynthesis (Figures 4C,D). Moreover, our results clearly showed that sediment re-suspension did not have substantial negative effects on the overall O\textsubscript{2} conditions within the water-column (Figures 4A,B) as previously suggested (Erftemeijer and Lewis, 2006), thus underpinning the critical importance of silt/clay leaf covers.

Plants with leaf silt/clay-cover exhibited internal meristematic tissue anoxia at higher water-column pO\textsubscript{2} levels (~45% of air equilibrium) than plants without leaf silt/clay-cover (~37% of air equilibrium), thus correlating with the lower passive O\textsubscript{2} influx into leaves with silt/clay-cover during night-time determined in the laboratory experiments (Figures 1,3). The silt/clay-induced negative effect on the intra-plant O\textsubscript{2} status was aggravated during prolonged exposure to fine sediment particles in the water-column (Figure 5), where the critical water-column O\textsubscript{2} level for Z. muelleri increased to ~63% air saturation after ~54 h of exposure to experimentally manipulated silt/clay re-suspension (Figure 5). Seagrass plants with leaf silt/clay-cover were thus more vulnerable to low water-column pO\textsubscript{2} at night-time and are exposed to an increased risk for H\textsubscript{2}S intrusion.

Proof of H\textsubscript{2}S intrusion in seagrasses has only been demonstrated in situ once (Borum et al., 2005) and never under conditions of such high water column pO\textsubscript{2} as in the silt/clay-treated plants of this study, which was in strong contrast to the control treatment, where no H\textsubscript{2}S intrusion was detected (Figures 4C,D). Anoxic conditions in the roots, rhizome and basal meristem of seagrasses lead to ceased radial O\textsubscript{2} loss (ROL) from the below-ground tissue into the immediate rhizosphere and thus resulted in sediment-produced H\textsubscript{2}S reaching the below-ground tissue surface (Brodersen et al., 2015b). If H\textsubscript{2}S enters the plant e.g., via, the root apical meristems, the transport of H\textsubscript{2}S to the basal leaf meristem is relatively fast as it occurs via gas-phase diffusion in the aerenchyma (Pedersen et al., 2004) and this may lead to chemical asphyxiation and thereby enhanced seagrass mortality (Lamers et al., 2013). Normally, H\textsubscript{2}S intrusion is prevented by plant-derived ROL creating oxic sediment microniches that are sustained as long as the below-ground tissue is supported with sufficient O\textsubscript{2} from the leaf canopy (Pedersen et al., 2004; Brodersen et al., 2015b, 2016). Mature regions of seagrass roots do not leak O\textsubscript{2}, but instead possess barriers to ROL, and thereby most likely to H\textsubscript{2}S intrusion, composed by Caspian-band like structures in the root endodermis (Barnabas, 1996; Enstone et al., 2003). This important anatomical cell-wall modification significantly reduces the consumption of O\textsubscript{2} along the internal diffusion path and thereby ensures an effective O\textsubscript{2} transport to the most distal parts of the seagrass plant (Colmer, 2003). At sunrise, photosynthetic O\textsubscript{2} evolution in the leaves of the silt/clay-treated plants lead to enhanced internal meristematic pO\textsubscript{2} and thereby re-oxidation of intruded H\textsubscript{2}S around 08:00–10:00 in the morning (Figure 4D), where after the H\textsubscript{2}S concentration remained below the detection limit.

Unfortunately, such in situ microsensor measurements are extremely challenging to obtain as positioning multiple microsensors simultaneously inside the tissue at the base of the shoot while “SCUBA diving” is very challenging and time consuming, and due to, e.g., sensor breakages during night-time as a result of fish foraging in the investigated seagrass meadow, as well as, time constrains such as daylight hours when positioning the sensors. Extreme changes in weather conditions did not allow us to perform additional replication. However, our results are very consistent with previous findings in situ and in the laboratory (e.g., Pedersen et al., 2004; Borum et al., 2005, 2006), showing H\textsubscript{2}S intrusion as soon as the aerenchymal tissue becomes completely anoxic, which only occurred for longer time periods in the silt/clay-treated plant during prolonged exposure to sediment re-suspension (Figures 4, 5). This clearly demonstrates that compromised photosynthesis as a result of prolonged exposure to sediment re-suspension and deposition of fine sediment particles on seagrass leaves can result in inadequate internal tissue aeration and thereby reduced below-ground tissue oxidation capacity, which leaves the plant exposed to intrusion of reduced chemical compounds such as H\textsubscript{2}S. The intra-plant O\textsubscript{2} conditions during night-time were similar in both the control plant and silt/clay-exposed plant during Series 1, whereas this changed completely during Series 2, where the same plants showed a very different response and the silt/clay-exposed plant became completely anoxic within a few hours after sunset at high water-column pO\textsubscript{2} simultaneously with the recording of rapid H\textsubscript{2}S intrusion (Figures 4, 5).

Settling of fine sediment particles onto seagrass leaves thus severely hampers the plants’ performance in both light and darkness, and thereby the health of the seagrass community as a whole. Silt/clay-induced compromised photosynthesis seemed to be the most important impediment to seagrass health in our study. Dredging-induced increased water turbidity therefore represents a severe threat to seagrass communities due to its adverse effects on internal O\textsubscript{2} status, and therefore can explain
the often major seagrass die-off events observed during excessive dredging activities (e.g., York et al., 2015), especially if carried out during summer-time where seagrasses are more prone to tissue anoxia owing to higher respiratory needs (Staehr and Borum, 2011; Raun and Borum, 2013); thus emphasizing the need for minimizing stress-inducing dredging operations for seagrass health.

In conclusion, the present study emphasizes the importance for seagrasses to maintain protective plant-derived oxic microshields within their rhizosphere, as sediment detoxification via ROL prevents H$_2$S from accumulating to very high toxic levels in the sediment and thus prevents H$_2$S from reaching the plant surface at the most vulnerable regions of the plants (Carlson et al., 1994; Brodersen et al., 2015b). Silt/clay-induced H$_2$S intrusion into Z. muelleri seemed tightly coupled to prolonged exposure to sediment re-suspension, such as typically found during harbor dredging activities (York et al., 2015) and resulting from river plumes (Petus et al., 2014). Leaf silt/clay-covers thus impeded the plants’ performance and thereby their resilience toward H$_2$S intrusion. This was as a result of a combined negative plant response to the reduced light availability for photosynthesis, thicker DBLs around leaves and enhanced leaf surface microbial respiration rates, all leading to inadequate internal aeration and reduced below-ground tissue oxidation capacity (Figure 4). Turbidity-generating activities such as dredging operations in close proximity to seagrass meadows can have strong negative effects on the fitness level and health of seagrasses through multiple pathways and may lead to increased seagrass mortality.

REFERENCES


AUTHOR CONTRIBUTIONS

KB, OP, MK, PR, and MR designed the research. KB, OP, KH, VS, and AF conducted the experiments. KB processed the data with help from OP and KH. KB, OP, MK analyzed the data. KB wrote the manuscript with editorial help from OP, MK, PR, and MR. All authors have given approval to the final version of the manuscript.

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SUPPLEMENTARY MATERIAL

The Supplementary Material for this article can be found online at: http://journal.frontiersin.org/article/10.3389/fpls.2017.00657/full#supplementary-material


Conflict of Interest Statement: The authors declare that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest.